

# Heavy Metals Contaminated Environments and the Road Map with Phytoremediation

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## Abstract

This review presents the justification of research into heavy metals and clean-up technologies with particular emphasis on phytoremediation of heavy metals contaminated soils. The generating sectors and heavy metal contaminant contributors are reviewed alongside likely types. The new paradigm by researchers in response to negative impacts of various clean-up methodologies has been discussed. This work reports limitations and prospects of phytoremediation in view of the future direction and reviewed issues with bioavailability, mobility, and response by plants to heavy metals in their environment. It reviews various plants used in phytoremediation of heavy metals and level of success recorded by teeming researchers. Diverse options available for optimization of this relatively novel technique to enhance performance have been elaborated. Suggestions for responsible abandonment of emitting sites and facilities, safety issues and appropriate disposal and management methods for plants used in this technology have been documented.

## Keywords

Phytoremediation, Heavy Metals, Soil Contamination, Biosafety, Phytotoxicity

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## 1. Introduction

The quest to meet man's basic need for food, clothing and shelter has continually heightened pollution of the biosphere, especially the soil. This in response has elicited keen interest among researchers to access, remediate (in a responsible manner), monitor and build industrial capacities through reliable data generated. Of contaminants identified, Alkorka [1] has labeled heavy metals as the most dangerous substances in the environment due to their recalcitrance, unlike organic counterparts, increased mobility under changing physical and chemical

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conditions and persistent toxicity to biota. The addition of excess common heavy metals with various human activities like arsenic, cadmium, lead, and zinc in zinc smelting operations [2] especially with unmuffled roasting furnaces, has heightened levels of these environmental terrors. A heavy metal is any metal or metalloid of environmental concern and has been defined giving great contributions from Dufus [3], Hawkes [4] and Blake [5], based on their density ( $>3.5 \text{ g/cm}^3$  to  $7 \text{ g/cm}^3$ ), atomic weight ( $>22.98$  for Na), atomic number ( $>20$  to 92), toxicity (toxic) or periodic table position (all the metals in Groups 3 to 16 that are in periods 4 and greater). There is no widely agreed definition of a heavy metal. Activities such as mining [6], smelting [7], oil exploration and exploitation [8], manufacturing facilities [9], illicit refuse disposal [10], electroplating, fertilisers [11] and paints manufacture, wood treatment, vehicular emissions especially tetraethyl lead [12], sewage sludge application to agricultural soils [13], firearm training giving off lead from lead azide or lead styphnate used in firearms [14], etc., have been implicated in heavy metals contamination of soil. Heavy metals enter plant, animal and human tissues via air inhalation, food, water and contact. Arsenic, which is low in nature although toxic [15] can become elevated as a by-product of Zn, Pb and Cu smelting activities [2]. Thus, arsenic is classified in EPA's Group A as a human carcinogen and is regulated as such. Carey *et al.* [16] describe the binding of AsIII (inorganic specie of As) to Cys residues as one that disrupts protein structure and function, thus affecting many key metabolic processes in the cell, such as oxidative phosphorylation, glutathione production, ATP synthesis, fatty acid metabolism, and gluconeogenic pathway. According to Lenntech [17], International Agency for Research on Cancer (IARC) has listed cobalt and cobalt compounds within Group 2B (agents which are possibly carcinogenic to humans). Cadmium and lead have got listed as EPA class B<sub>1</sub> metals and are very toxic [15] at high concentrations. Alkorta [1] describes Cd as highly mobile in plant-soil systems, although with less evident toxicity, and is described as by-product of Zn and Pb mining and smelting. Lead is more complex to remove especially when introduced to soil matrix but they can be mobilized into the solution phase by changing the soil pH, temperature, redox potential, and soil organic matter composition [18]. Zinc, though not classified as mutagenic or carcinogenic [15], is toxic at high concentrations and constitutes the most mobile heavy metal because it is present as soluble compounds at neutral and acidic pH values [19]. People who live near hazardous waste sites, nuclear power plants and mines, work in the phosphate industry, eat produce from contaminated soil, or drink water from a uranium waste disposal point may experience a higher uranium exposure than other people. Although, root vegetables (such as radishes) may contain higher than usual concentrations of uranium (as heavy metal and as a radiological hazard) as plants absorbing uranium through their roots and store it there, they are removed when the vegetables are washed [15].

Lenntech [15] also reported higher-than-usual exposure for artists that use them for glasswork, despite the fact that uranium glazes are banned. While uranium itself is not particularly dangerous, some of its decay products do pose a threat, especially radon, which can build up in confined spaces such as basements [15]. However, Mclay [20] reported toxic effect on the living cells as processes of carbohydrates metabolism are inhibited by the inhibition of enzyme systems (associated with hexokinase at the sites of ATP surface-building through magnesium-hexokinase mechanism), and suggests this could account for why many people are gaining so much weight in the last couple of years. Early studies on the biological effects of uranium showed that uranium salts given by mouth presented a hazard as a mild poison causing death.

Detrimental effects of heavy metals can be viewed from their interference to the proper functioning of vital cellular components, such as structural proteins, enzymes, and nucleic acids, when they bind to them, although it may be symptomatic or not depending on type and dose. Important manifestations has been reported as shown in **Table 1** to underscore the importance of assessment, management and prevention of heavy metals contamination. This review evaluates the state of heavy metals contamination and clean-up vis-a-vis phytoremediation to examine prospects ahead.

## 2. Overview of Plants Used for Phytoremediation and Recorded Degree of Success

### 2.1. Justification and Limitation of Technique

Heavy metals have impacted on the ecosystem through discharges as effluents, dust and/or leachate. Although many metals are essential, all metals are toxic at elevated concentrations, because they form free radicals thereby causing oxidative stress and can replace essential metals in pigments or enzymes disrupting their function [21]. In a bid to clean contaminated soils, different agencies, companies and researchers have employed leaving the contamination as it is and restricting the utilization of the land (Stegmann, 2001), complete or partial encap-

**Table 1.** Varying exposure conditions with common heavy metals in humans.

Element	Acute exposure	Chronic exposure	Condition
Cadmium	Pneumonitis (lung inflammation)	Lung cancer Osteomalacia (softening of bones) Proteinuria (excess protein in urine; possible kidney damage)	Cadmium poisoning
Mercury	Diarrhea Fever Vomiting	Stomatitis (inflammation of gums and mouth) Nausea Nephrotic syndrome (nonspecific kidney disorder) Neurasthenia (neurotic disorder) Parageusia (metallic taste) Pink Disease (pain and pink discoloration of hands and feet) tremor	Mercury poisoning
Lead	Encephalopathy (brain dysfunction) Nausea Vomiting	Anemia Encephalopathy Foot drop/wrist drop (palsy) Nephropathy (kidney disease)	Lead poisoning
Chromium	Gastrointestinal hemorrhage (bleeding) Hemolysis (red blood cell destruction) Acute renal failure	Pulmonary fibrosis (lung scarring) Lung cancer	Chromium toxicity
Arsenic	Nausea Vomiting Diarrhea Encephalopathy Multi-organ effects Arrhythmia Painful neuropathy	Diabetes Hypopigmentation/Hyperkeratosis Cancer	Arsenic poisoning
Zinc		Acute gastrointestinal and respiratory damages and acute heart, brain and kidney damages	Zinc toxicity
Nickel	Respiratory, skin and heart aches	Total defects in infants, cardiovascular defects and defects of the musculoskeletal system	Nickel toxicity
Copper	Flu-like condition known as metal fever, vomiting, diarrhea, stomach cramps, and nausea	Wilson's disease, characterized by a hepatic cirrhosis, brain damage, demyelization, renal disease, and copper deposition in the cornea, and even death	Copper poisoning
Selenium	Decreased synthesis of thyroid hormones and metabolism of growth hormone and insulin-like growth factor-1	Neurotoxicity, nail and hair loss and dermatitis	Selenium poisoning
Cobalt	Vomiting and nausea, vision problems	Lung effects, such as asthma and pneumonia, heart problems, thyroid damage	Cobalt Poisoning
Uranium	Mild aches	Kidney disease, cancer with enriched uranium, inhibit the processes carbohydrates metabolism by the inhibition of enzyme systems, childhood and breast cancer in high prone zones	Uranium Poisoning

Afal & Wiener [69]; Chashschina *et al.* [70]; Lenntech [17]; Vinceti *et al.* [71].

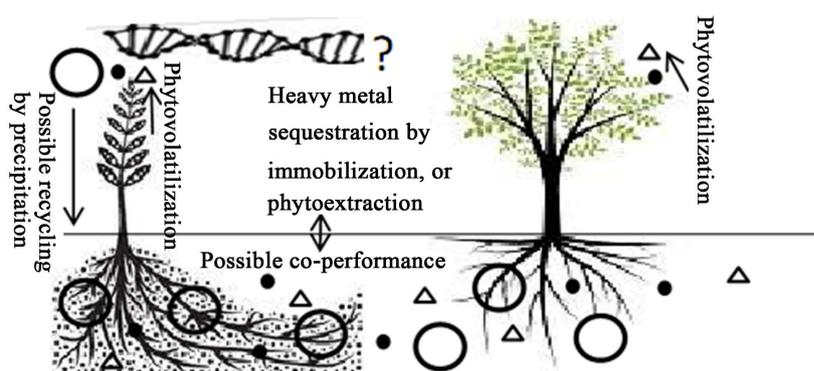
sulation (isolation and containment) of the contaminated site, dig-and-dump, burying of contaminated soil, dilution of contaminated soil with clean soil, electrokinetics [22] for low permeable soils, use of chemicals (oxidizing agents, acids and solvents) [22] and dispersants (*in situ* or *ex situ*), and thermal treatments (*ex situ*), potentially endangering biota [23] and causing deterioration of groundwater quality. Conventional methods of remediation have been estimated at \$10 to 1000 per cubic meter besides its environmentally destructiveness [24] while phytoextraction costs are estimated to be as low as \$0.05 per cubic meter [25]. To achieve efficiency, and cost-effectiveness, biological methodologies with environmental compatibility, have evolved with bioremediation (with bacteria, fungi, algae, plankton, and protozoa) serving well in clean-up of organics [13] as well as stabilization of heavy metals but not their breakdown. An emerging technology, phytoremediation uses plants and their associated microbes for the removal of pollutants from the environment or to reduce their toxicity [26]. Several advantages ranging from aesthetics purposes, cost-effectiveness, less environmental disturbance, less

technical—know how on implementation, have been associated with phytoremediation of heavy metal polluted sites. Raskin and Ensley [27] reported the shielding advantage, as ground cover of plants, preventing the blowing of contaminated dust around the neighborhood apart from the widely acclaimed metal hyperaccumulating property of the plants. Despite identified limitations as enumerated by Pilon-Smits [28], which include handling concerns of phytoaccumulators after clean-up, limited applicability of this method to a heavily contaminated soil, given long time required for cleaning up the contaminated site, limitation of technique to bioavailable fraction of pollutant in the soil, phytoremediation has enjoyed wide studies among researchers. In situ design has presented lowered cost and impact on the ecosystem. Overall, appropriate remediation technique for a heavy metal contaminated site is time consuming, site specific and tricky and therefore needs full understanding of available options, nature of metal, accessibility to site and available resources. Among identified phytoremediation classifications, phytoextraction, phytostabilization, phytovolatilization (for Hg and Se) [28], and rhizofiltration are most implicated in the remediation of heavy metals contaminated environments. Different plants, given their local advantages may fit a case and not the other in phytoremediation of heavy metals in soils because interest heavy metals are present in soil in different fractions. They can therefore be dissolved in soil solution (Figure 1), attached to exchange sites on inorganic soil constituents, adsorbed to inorganic soil constituent, attached to insoluble organic matter, or as precipitates of pure or mixed solids, based on the properties of the individual metals.

## 2.2. Brief Classifications of Phytoremediation

### 2.2.1. Phytoextraction

This is a technique that concentrates contaminants in the harvestable parts of plants capable of phytoaccumulating high biomass production. It has been reported to preserve structure and fertility of such soils as less disturbance is required [29]. Hyperaccumulators are plants that contain greater than or up to 0.1% *i.e.* more than (1000 mg/g) of copper, cadmium, chromium, lead, nickel cobalt or 1% (>10,000 mg/g) of zinc or manganese in the dry matter and their cropping may be repeated until desired result is achieved. For cadmium and other rare metals, it is >0.01% by dry weight [30]. Their phytoextraction however, may be limited to 3 feet and 10 feet [25] from surface for soil and ground water heavy metal contamination respectively. Areas of high metal contamination may signal potential hyperaccumulator species. Baker and Brooks [30] reported prevalence of various metal hyperaccumulation in the Brassicaceae family as 87 species (of about 400 from 22 families) from 11 genera has been documented. Brooks *et al.* [31] posits that if phytoextraction could be combined with biomass generation and its commercial utilization as an energy source, then it can be turned into profit making operation and the remaining ash can be used as bio-ore and this forms the basic principle of phytomining.



**Figure 1.** Mechanism and current status of phytoremediation of heavy metals in contaminated soils metals are sequestered in soil by immobilization, transformation and uptake into plant tissues. Horizontal line shows harvestable plant part that is mostly ashed. Metals like Se and Hg have been shown to be successfully volatilized in non toxic form. However, possibility of recycling by precipitation and leaf drop is feared. Co-plant rhizosperic interactions have been shown to modify performance of species at metal uptake as conditions in the shared rhizospheres may influence metal bioavailability to neighboring plants. Screening of plants for multi-metal hyperaccumulators is ongoing considering competition by certain heavy metals such as Cu and Zn, Ni and Cd for same membrane protein transporters. Figure also shows the future of this technique with genetic engineering in root, leaf, biomass, tolerance, accumulation potentials of identified species. Note different heavy metal contaminants (in circles and triangles) undergoing either immobilization or mineralization via natural plant exudation or supplementation with synthetic chelators for possible tissue uptake.

### 2.2.2. Phytostabilization

This is a complexation technique that dwindles contaminant mobility and bioavailability in soils, sludges and sediments thereby limiting biomagnification via erosion and leaching. This localised contaminants however may require some level of monitoring. Locally available materials like crushed mussel shell may be explored as reported by Garrido-Rodríguez *et al.* [32] who observed diminished copper desorption and mobilization rates in copper enriched vineyard and mine samples even at low pH of 3.

### 2.2.3. Phytovolatilization (for Hg and Se)

This technique involves the venting of contaminants or their metabolites via the leaves (**Figure 1**) to the atmosphere. Pilon-Smits [28] enlisted some inorganics (Hg and Se) capable of existing in volatile forms and volatile organic carbons as susceptible to this technique. Meagher *et al.* [33] reported absorption of Hg(II) and volatilization of Hg(0) by *N. tabacum* and an engineered model *A. thaliana* (with a gene for mercuric reductase). Similarly transformed yellow poplar (*Liriodendron tulipifera*) plantlets showed resistance to, and grew well in Hg contaminated regimes. Successes have also been recorded with tritium removal [34] with phytovolatilization. Utmost care however needs to be taken to exclude possibility of recycling by precipitation (**Figure 1**) when considering a contaminant metabolic route thereof.

### 2.2.4. Rhizofiltration

This is the *in situ* or *ex situ* adsorption, absorption, and precipitation of inorganic and organic contaminants from mild contaminated environments, especially liquid discharges, using plant (terrestrial and aquatic) roots. This is most ideal for metal (Ni, U, Cr, Cd, Zn, Pb and Cu) contaminants that are basically retained in the root section [29] [35] [36]. Raskin and Ensley [27] suggested a preferential use of terrestrial over aquatic plants since they have a fibrous and much longer root system, increasing the amount of root area.

### 2.2.5. Phytotransformation

This is the total or partial degradation of organic contaminants by breakdown or transformation into simpler forms that are incorporated into plant tissues. This breakdown may be by plant enzyme (usually dehalogenases, oxygenases and reductases) [37] or rhizodegradation (breakdown of organic contaminants usually as fuel or solvent, by rhizospheric microbial activities).

### 2.2.6. Useful Ratios for Interpretation

To further qualify performance, some ratios like concentration ratio (the ratio of the metal concentration in the shoot of the plant to that in the soil based on wet weight) has been used to indicate presence or absence of accumulation in uptake studies. Similarly, bioconcentration factor also known as bioaccumulation factor or enrichment coefficient [9] is often used and is computed as the ratio of a given metal concentration in the plant tissues (dry weight) [38] at harvest to the concentration of the metal in the concerned environmental component. Larger values are taken to describe better phytoaccumulation capability. A phytoextraction coefficient of 1.7 was reported for *Brassica juncea* and it has been found that a lead concentration of 500 mg/l is not phytotoxic to *Brassica* species [21]. Another ratio, translocation factor has been computed as the ratio of the metal concentration in root tissues to their counterparts in shoot tissues [39]. Calculation of the time required for cleaning up the soil with plant can also be determined using the amount of metal accumulated in the harvestable parts of the plant (shoot), and the bioavailable metal present in the soil.

### 2.2.7. Recorded Success

Phytoremediation of heavy metal contaminated soil has attracted scientists of diverse origin, culture, race and disciplines because the environment is involved. The heightened interest is not restricted to researchers but also to industries and most importantly the government nationally and internationally springing up relevant study grants among others, as the global thirst for cheaper, simpler and more eco-friendly technologies rockets. Exposure responses, removal efficiencies and phytoextraction and accumulation possibilities have been explored by several researchers. It is known that to enhance metal solubility, plants either excrete organic ligands or lower the soil pH in the rhizosphere. To mimicksuch natural enhanced metal solubility via exudation or pH reduction. Romkens *et al.* [40] reported success with administration of synthetic chelates such as ethylenediaminetetraacetic acid (EDTA), nitrilotriacetic acid, pyridine-2-6-dicarboxylic acid, citrate, nitric acid, malate, histidine [41],

hydrochloric acid and fluorosilicic acid at phytoremediation studies. Many chelates have been tested and the order of effectiveness in increasing Pb desorption from the soil was EDTA > Hydroxyethylethylene-diaminetriacetic acid (HEDTA) > Diethylenetriaminepentaacetic acid (DTPA) > Ethylenediamine di(o-hydroxyphenylacetic acid) EDDHA [42]. The addition of lime or organic matter has been reported to lower heavy metal metal solubility [28] at near phytotoxic levels.

Also, Mathew [7] reported reduced % bioavailability with increased contaminant load for As (28.8, 2.1 and 0.26 mg·kg<sup>-1</sup>) and Cd (26.2, 15.2 and 9 mg·kg<sup>-1</sup>) for low, medium and heavy contaminated smelting site respectively. Observed trend was reversed with zinc while no particular order was established by the researcher for Pb. This pattern manifested in higher shoot uptake level for Zn in corn as reported Mathew [7]. Although only the bioavailable fraction of heavy metal in the soil is subject to phytoremediation, some of the tightly bound heavy metals can become bioavailable with exhaustion of the bioavailable fraction. Arsenic was higher in roots than in leaves for corn [7] but higher in leaves with Chinese Brake ferns [43]. Corn gave a better heavy metal uptake when compared to sunflower although both gave significant successes as observed by Mathew [7] and Spirochova *et al.* [44].

Mechanism of heavy metal accumulation (Figure 1) is being studied. Cosio *et al.* [45] investigated the mechanisms of Zn and Cd accumulation in different plant species (*T. caerulescens* “Ganges” and *A. halleri*) through ion compartmentation by measuring their short term <sup>109</sup>Cd and <sup>65</sup>Zn uptake in mesophyll protoplast and suggested a regulation mechanism in place. Response of plant parts to exposure is under wide research. In that vein, Nwaichi *et al.* [46] and Whiting *et al.* [47] found increased root biomass and root length in favour of hydrocarbon and heavy metal phytotoxicity in *V. subterranean* and *T. caerulescens* respectively. Co-plant rhizospheric interactions have been shown to modify performance of species at metal uptake. In their study, Gove *et al.* [48] observed up to 2.4 factor increase for Cd uptake in *H. vulgare* plants when grown alongside (with no barrier) of *T. caerulescens*. The case however, was reversed (decrease) for Zn uptake in *H. vulgare*. This may have arisen due to alterations in conditions in the shared rhizospheres and which may have influenced metal bioavailability to neighboring plants. Similar observations were made by Wenzel *et al.* [49] using *S. alfredii* and *Z. mays*. Current methods for recovery of heavy metals from plant biomass of hyperaccumulators are unclear given increasing cost with energy used for gasification. Further, pollution could result if metals are volatile giving ashing (of metal-rich biomass) and recovery methods, where disposal as hazardous waste is not followed. Competition by certain heavy metals such as Cu and Zn, Ni and Cd for same membrane protein transporters was reported by Clarkson and Luttge [50]. This could have a serious implication in the choice of plants for the clean-up in co-contamination systems.

Inorganic and organic agents, including EDTA, citric acid, elemental sulfur or ammonium sulfate [51] [52], and urea have been applied to soils to improve phytoextraction potentials of different plant species. These act as chelating material in addition to natural plant exudates. The importance of the activities of rhizosphere-associated microbes (such as several strains of bacillus and pseudomonas) at degradation of organic pollutants has been proven helpful in the *B. juncea* phytoextraction of the Cd [41]. Molecularly, enhancement of expression of protein transporters into the root and shoot and enzymes that could modify and conjugate metals as well as enhanced level of root, xylem and phloem chelators (NA, GSH, acids) have been reported by Dhanker *et al.* [53]. Possible upregulation of degrading enzymes from roots has a great potential for the secretion of compounds that stimulate microbial density or activity as in rhizodegradation [53]. Genes encoding plant arsenate reductases for example have been isolated and characterized from Arabidopsis [54] and many other species including rice, *Holcus lanatus*, and *P. Vittata*. Co-expression of both γ-ECS and PCS in Arabidopsis gave a greater effect on As accumulation and tolerance than over-expression of either gene alone [55]—this firmed up preferential successes with combined approach. Transgenic processes will therefore offer a lot to the field of phytoremediation in terms of relevant novel gene introduction and/ or modification of existing types.

In all, most of the Species used successfully in phytoremediation include Corn and Sunflower [7] for Pb (25,008 mg·kg<sup>-1</sup>), Zn (94,420 mg·kg<sup>-1</sup>), As (1658 mg·kg<sup>-1</sup>) and Cd (1281 mg·kg<sup>-1</sup>); Tossa jute (*Corchorus olitorius*) for As and Cr [56]; Chinese brake fern [57] for As and not Zn; Indian mustard (*Brassica juncea*) in soils up to 200 mg·kg<sup>-1</sup> [58], willow clones (*Salix*), alpine penny-cress (*Thlaspi caerulescens*) up to 390 mg·kg<sup>-1</sup> [59], sunflower (*Helianthus annuus*) and corn (*Zea mays*) (up to 90 mg·kg<sup>-1</sup> [44] for Cd; *Brassica juncea* up to 500 mg·kg<sup>-1</sup> and 1500 mg·kg<sup>-1</sup> [7] *Helianthus annuus* and *Zea mays* up to 16,000 mg·kg<sup>-1</sup> [44], *Piptatherum miliaceum* (Smilo grass) up to 1550 mg·kg<sup>-1</sup> [60], *Thlaspi praecox* up to 67,940 mg·kg<sup>-1</sup> [61], *Hemidesmus indicus* up to 65% of 10,000 mg·kg<sup>-1</sup> [62] for Pb; *H. annuus* up to 350 ppb for uranium (95% reduction to 5 ppb in 24 h)

[63], *Piptatherum miliaceum* (Smilo grass) up to  $600 \text{ mg}\cdot\text{kg}^{-1}$  [60], *Pteris vitatta* (a fern) up to  $14,500 \text{ mg}\cdot\text{kg}^{-1}$  for As without signs of toxicity [44], *A. serpyllifolium* for up to Cr ( $283 \text{ mg/kg}$ ), Cu ( $264 \text{ mg/kg}$ ), Pb ( $1433 \text{ mg/kg}$ ) and Zn ( $377 \text{ mg/kg}$ ) [64], *Helianthus annuus* and *Zea mays* up to  $75,000 \text{ mg}\cdot\text{kg}^{-1}$  [44]) and *Thlaspi caerulescens* up to  $3259 \text{ mg}\cdot\text{kg}^{-1}$  [65] and  $26,000 \text{ mg}\cdot\text{kg}^{-1}$  for Zn and up to 22% of soil exchangeable Cd from contaminated site [43] [66] [67] found 95% of As localised in the harvestable portion of fern, with higher amount in older fronds, while Salt [41] observed the reverse situation in Indian Mustard for As. Similar pattern of higher accumulation on older leaves was observed in *Typha domingensis* by Hegazy *et al.* [9] in industrial wastewater remediation. Linear correlation between shoot concentrations and soil concentrations of heavy metals could be interpreted as confirmatory analysis for phytoremediation of metal using such specie [61].

### 3. Prospects of Phytoremediation of Heavy Metal Contaminated Soils

Phytoaccumulators could be subjected to metal recovery as a decontamination approach after compaction in a responsible manner and clean biomass can support agricultural management practices. Induced phytoextraction or chelate assisted phytoextraction is a way to go, to release most metals and improve phytoextraction of heavy metals. This will require a more comprehensive comparative studies of available chelators for scaled and holistic performance. Molecular biology (with streamlined strategies for monitoring different stages of genetic manipulation) could be employed to improve required traits such as dense rooting system, high growth rate, disease resistance, selectivity to metals, high resistance to toxicity, improved biomass production, enhanced accumulator genes, etc. to enhance patronage of technique. Molecular identification of metal specific transporter genes for toxic heavy metal species are most important for those serving extra position as plant nutrient so as to avert starvation while toxic species are extracted. Broader anatomical and physiological studies of screened and identified plants for phytoremediation could open up a new area for genetic engineers in this field. Invasive species however, may not fit in these genetic manipulations as spread may be uncontrolled. Furthermore, more field trials should be done with screened species as behavioural differences in hydroponic conditions, pots and real world may suggest potential differences.

Synergistic toxic effects of multiple heavy metal contamination may require improved variety. For example, Islam *et al.* [56] reported that high level ( $100 \text{ mg}\cdot\text{kg}^{-1}$ ) As plus Cr caused a further decreased plant growth and chlorophyll content, increased MDA and  $\text{H}_2\text{O}_2$  contents as well as antioxidative enzymes activities significantly ( $P \leq 0.05$ ) and less severe inhibition of plant growth and oxidative damage was observed in O-795 (Cr-tolerant jute) than in O-9897 (Cr-sensitive jute) indicating variety O-795 had more efficient defense system to mitigate heavy metal induced oxidative stress. Some plants (such as *Sesuvium portulacastrum* L.) however possess the molecular and physiological flexibility [68] to deal with such parched sites. Also chelating agents like synthetic ethylenediamine tetra acetic acid (EDTA) was applied to Pb contaminated soil (total soil Pb  $2500 \text{ mg}\cdot\text{kg}^{-1}$ ) and this increased the amount of bioavailable lead in the soil and caused a greater accumulation in plants (*Zea mays* (corn) and *Pisum sativum* (pea)) from less than  $500 \text{ mg}\cdot\text{kg}^{-1}$  to more than  $10,000 \text{ mg}\cdot\text{kg}^{-1}$  [42]. This can therefore be leveraged upon.

Most researchers have made choices of their plants species based on literature, climatic conditions of study areas, specie availability and density, growth and harvest edges [19], biomass yield and tolerance [10] [28], etc. Biotechnology techniques however is currently used to develop plants with even better characteristics for phytoremediation such as ability to accumulate multiple metals. However, molecular mechanisms of heavy metal detoxification and tolerance in identified plants needs a deeper understanding.

Selection of high biomass weeds (non-edible, disease resistant and tolerant plants) to restrict the biomagnification of heavy metals into the food chain may advance the viability of phytoremediation (especially of phytoextraction) and may have implications in renewable energy and biodiversity preservation.

### 4. Conclusion

Heavy metals can accumulate in organisms as they are hard to metabolize. Researchers over the last ten years have globally patronised by phytoremediation studies to tackle heavy metal removal from contaminated environments although with a load of concerns over biomass management and the pace of the technique. This environmentally friendly and relatively cheap process is fast emerging as a viable alternative to various conventional remediation methods. In a developing country like Nigeria, where most identified sources of contamination are in the boom, in a bid to meet the need of its teeming population, phytoremediation will be a good fit. Since soil

clean-up entails return of soil to a state where it can perform its ecological functions including establishment of biota communities, it supports prior to disturbance, assessment of soil community shifts and their physiological profiles should be done to complement physical and chemical data on abandonment of site. Government and relevant agencies should go beyond paper at ensuring compliance to set regularly reviewed standards to protect and reclaim our soils and rendering measurable support to researchers in this area, while creating awareness. A review of the status of phytoremediation as a technology is timely to equip researchers and policy makers with gaps, successes and potentials embedded in this novel technology in managing the heinous environmental plaque of heavy metal contamination.

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