

Hemin Based Biomimetic Oxidative Degradation of Acid Orange 7

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Abstract

Degradation of dyes is an important environmental issue. In order to avoid the carcinogenic risks in anaerobic-aerobic biological process for wastewater containing azo dyes, a hemin based biomimetic oxidative degradation of azo dyes was developed. Acid orange 7 (AO7) was selected as the model for azo dye and the high efficient degradation was achieved in hemin/H₂O₂ system at pH 11.0. Degradation could be described by a pseudo-first-order kinetic model. The order of dependence on H₂O₂ concentration was significantly larger than that of hemin. Coexisting anions sulphate and chloride had little effect on the degradation, but reductive sulphite dramatically inhibited the degradation. The protic solvent 2-prophanol obviously promoted the degradation. Azo chromogenic group was destroyed quickly and some smaller intermediates formed. Active species oxoferryl porphyrin π -cation radical '+PFe^{IV}=O generated from heterolytic cleavage of O-O in H₂O₂ catalyzed by hemin play the main roles in degradation and reaction pathways were proposed.

Keywords

Biomimetic Oxidative Degradation, Azo Dyes, Hemin

1. Introduction

The removal of dyes from water is one of the most important issues for industries such as textile, paper, leather, pharmaceutical, food and so on [1]-[5]. Azo dyes represent 50% - 70% of all dyes and 10% - 15% are released into wastewater. Most wastewater containing azo dyes is treated in wastewater treatment plants by traditional biological methods based on anaerobic-aerobic process [6]. These processes are relatively ineffective in degradation of dyes and significant parts of azo dyes release carcinogenic aromatic amines by microbial reductive cleavage of -N=N- under anaerobic condition and toxicity increases instead [7] [8]. So it is necessary for azo dyes to become smaller and less toxic molecules by destroying the azo chromogenic groups before draining.

Oxidative degradation of azo linkage is approved to be more effective than anaerobic reductive cleavage and some oxidative systems have been reported [6] [9]. However, there are some disadvantages in chemical oxidative

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systems [4] [6] [9]-[11], such as narrow and acid pH range, long reaction time, use of non-convenient chemical reagents and generated large amount of sludge. Most wastewater from textile effluents are highly alkaline [12] [13], so it is cost-prohibitive because of large amount of acidic and alkaline chemicals used for pH adjustment.

The effective biological systems for oxidative degradation of azo dyes mainly are peroxidase systems, such as horseradish peroxidase/H₂O₂ [14] [15], manganese peroxidase/H₂O₂ [16] and lignin peroxidase/H₂O₂ [17]. The group ferriprotoporphyrin IX in the active center of the enzymes plays the main catalytic role in peroxidase systems. Hemin is biomimetic compounds of peroxidase and have shown very good efficiency for hydrogen peroxide activation in alkene epoxidations [18] and in degradation of dyes [3] [5] [12] [19]. There were different characteristics and mechanisms reported about the degradation of dyes catalyzed by hemin. So it is necessary to understand the catalytic reaction of hemin on the oxidative degradation of dyes deeply. Therefore acid orange 7 (AO7) was selected as a model for azo dye [3] and efficient oxidation system with low concentration of hydrogen peroxidative degradation of AO7 were explored. Then the reaction kinetics was studied and the kinetic rate coefficients were determined based on the experimental data. Based on the kinetic and characteristic analysis of degradation process, the possible degrading pathway was proposed.

2. Experimental

AO7 was dissolved in 0.2 M Na₂HPO₄-NaH₂PO₄-NaOH buffers for desired pH stock solutions with 0.67 mM. Hemin was dissolved in the N,N-Dimethylformamide with 0.1 mM and diluted 10 times with water before used.

The experiments were conducted by varying the parameters such as pH, hemin concentration, H_2O_2 concentration, temperature, co-exiting anions and protic solvent. For degradation experiment, the AO7 stock solution was diluted to the desired concentration with buffer solution with same pH, then corresponding volume of H_2O_2 and hemin solution were added in turn and the mixture reacted in a constant temperature bath. The related parameter values were shown in figure captions.

At certain reaction time, absorbance of samples at 484 nm was recorded with a T6 Uv-vis spectrophotometer (Purkinje Co., China) and the residual concentrations of AO7 was calculated. The buffer solution with same pH

was used to dilute immediately if necessary. The degradation ratio was evaluated as $\left(1-\frac{C}{C_0}\right) \times 100\%$, and C_0

and C were the initial and the residual concentration of AO7 respectively. Blank was done at the same time in each group experiments, in which the same volume buffer solution was used to take the place of hemin solution.

For FTIR and GC-MS analysis, the reaction mixture was acidified with 0.4 mL 1M HCl and kept stirring for 5 min with 2 mL CH₂Cl₂ for extraction of reaction products at certain time. Then the extracted solution was dehydrated with anhydrous sodium sulfate and concentrated to 0.5 mL. The FTIR spectra were recorded on a thermo Nicolet iS10 FTIR spectrometer. 10 μ L concentrated extract was deposited on a pure KBr thin disk and solvent was removed under the IR lamp. For GC-MS (Agilent 7890A/5975C), a HP-5 MS capillary column (30 m × 0.25 mm × 0.25 μ m) was employed for GC separation. The GC was operated in a temperature programmed mode with an initial temperature of 60°C held for 2 min, then ramped to 250°C with a 20°C/min rate.

The dissolved oxygen concentration profiles were recorded with a Dissolved Oxygen Meter (Mettler-Toledo SG6-ELK) at 15°C with magnetic stirring.

3. Results and Discussion

3.1. Degradation Kinetics Analysis

pH is a highly important factor for effective oxidation of AO7 in hemin/ H_2O_2 system. Degradation of AO7 achieved in alkaline medium and increasing degradation ratios and rate were observed with increased original pHs in our study. The similar results were reported when hematin, the hydorxylated hemin, was used to catalyze the conversion of orange II at higher peroxide concentrations [3]. Alkaline medium are benefit for reaction of hemin and AO7. De Villiers *et al.* [20] observed a maximum of the dimerization constant of ferriprotoporphyrin IX in aqueous solution at pH 7, decreasing afterward, so increased pH may be associated with disaggregation of hemin dimmers, which account for higher availability to substrate attack. On the other hand, AO7 is in favor of equilibrium shift to azo form for AO7 [1]. A pKa corresponding to the -OH of AO7 is 10.7 [21], so –OH can dissociate to form azo body with phenolate ion at pH 11.0 [22] and phenolate ion is more easily oxidized [23]

[24]. Negatively charged species of AO7 at higher pH may favor disaggregation while molecular of AO7 aggregates in neutral solution [3] [25]. The result is quite valuable because of highly alkaline of much textile wastewater [12] [13], and the hemin/H₂O₂ system can be used directly without pH adjusting.

With regard to the actual pH of dyeing wastewater and characteristics of AO7, the next study was carried out at pH 11.0. The degradation profile of AO7 follows an exponential pattern and can be described well according

to the pseudo-first-order kinetics reaction model, $\ln\left(\frac{C}{C_0}\right) = -k_{obs}t$, in which k_{obs} represents the reaction rate

coefficient and t represents the time. kobs is represented by the Arrhenius equation in linear form:

$$\ln k_{obs} = \ln k' + \left(-\frac{E_a}{RT}\right)$$
(1)

where k' is the frequency factor, E_a the activation energy, T the temperature and R the gas constant (8.314 J·mol⁻¹·K⁻¹). At constant pH, the frequency factor k' in Equation (1) is a function of the hemin and H₂O₂ concentration and they can be related by the following equation:

$$\mathbf{k}' = \mathbf{k}'' \left[\text{Hemin} \right]^p \left[\mathbf{H}_2 \mathbf{O}_2 \right]^q \tag{2}$$

in which [Hemin] and $[H_2O_2]$ are the concentrations of hemin (μ M) and H_2O_2 (mM) respectively,p and q the orders of dependence respectively. The order of dependence on concentrations of hemin and H_2O_2 at pH 11.0 is 1.190 and 1.596 respectively (see **Figure 1**) according to the experiment results.

The order of dependence on H_2O_2 is significantly larger than that of hemin and the result is in keeping with the highly sensitive detection for H_2O_2 of linked hemin on the PEI-AuNP in our study [26].

The increase of temperature results the increase of degradation rate, especially the initial 5 min. So the initial slope of the observed discoloration was used to calculate the rates at different temperatures and **Figure 1(c)** gives the Arrhenius plot of lnk_{obs} versus 1/T at constant H_2O_2 and hemin concentrations. According to the plot, E_a is about 37.0 kJ·mol⁻¹, which is lower than the Ea value (43.5 kJ·mol⁻¹) of AO7 oxidized by heterogeneous photo-Fenton system [27]. The E_a value suggests ion-molecule or radical-molecule reactions, which requiring activation energy [27] happens during the degradation of AO7.

The last degradation ratios are similar when temperature is above 15° C in our study, so temperature was not so important as other factors and the reaction can be carried out at room temperature. From this point, he-min/H₂O₂ system is very suitable for pre-treating the actual dyeing wastewater under different temperature.

3.2. Effects of Coexisting Anion

Chloride and sulfate sodium salts are frequently used dye-assisting chemicals in the dyeing process [28], and $SO_3^{2^-}$ is a kind of reductive anions. Their effects on degradation of AO7 are tested. The results in **Figure 2** show 10 mM $SO_4^{2^-}$ and Cl have little effect on the degradation of AO7, but 10 mM $SO_3^{2^-}$ dramatically inhibits the degradation. And the more $SO_3^{2^-}$ in the mixture, the more dramatic inhibition effects. So the reductive



Figure 1. Plots of \ln_{kobs} versus $\ln[\text{Hemin}]$ (a), $\ln[\text{H}_2\text{O}_2]$ (b) and 1/T (c) ([AO7] 0.18 mM). (a) $[\text{H}_2\text{O}_2]$ 0.70 mM, T 15°C; (b) [Hemin] 0.18 μ M, T 15°C; (c) [Hemin] 0.27 μ M, $[\text{H}_2\text{O}_2]$ 0.35 mM.



Figure 2. Effects of chloride, sulfite or sulfate ions on color removal of AO7 (a) and Effects of the sulfite concentration on the color removal of AO7 (b). Experiment conditions: [AO] 0.07 mM, $[H_2O_2]$ 0.21 mM, [Hemin] 0.09 μ M, T 15°C.

interfering anions reduce the degradation rate and ratio of AO7 because of the competition consumption of oxidant.

3.3. Characterization of Degradation Process

UV-vis spectra during degradation processes are shown in **Figure 3** and the peaks at 484 nm and 430 nm are correspond to the $n \rightarrow \pi^*$ transition of the azo and hydrazone form, respectively. Other two bands at 250 and 310 nm are attributed to the $\pi \rightarrow \pi^*$ transition of benzene and naphthalene ring, respectively [1] [2] [29] [30]. The main band at 484 nm decreased quickly, but there was a new band appeared obviously at 343 nm and increased at first 5 min, then decreased with time prolonged. At the same time, the absorbance at 250 - 255 nm also increased and a new should peak appeared at 252 nm. The inset of **Figure 3** is the absorbance evolution at 343 and 252 nm. These changes indicate that some naphthalene- and/or benzene-type intermediates form during degradation process.

Figure 4 presents the FTIR spectra of samples during degradation of AO7. In spectra of 0 min, the band at 1266 cm⁻¹ is assigned to the vibration stretching mode v(C–N) [31], and the band at 1384 cm⁻¹ is v(N=N) and o-quinone structure[32]. The bands near 1730 cm⁻¹ is attributed to the v(C=O) and 1459 cm⁻¹ is v(C=C) in aromatic ring (benzene or naphthalene ring). During oxidative process, the band at 1266 cm⁻¹ became weaker and soon disappeared, which was in accord with the peak changes at 484 nm in UV-vis spectra. This indicates the break of bands C-N in the dye molecular. The new bands near 1723 cm⁻¹ was in accord with peak changes at 343 nm in UV-vis spectra and could be associated to the aromatic rings change.

Figure 5 shows the dissolved oxygen(DO) evolution during AO7 degradation. It was observed the DO increased at first, then decreased with time. It suggests that the reaction of AO7 in hemin/system at selected conditions involves O_2 releasing. Kalyanararna *et al.* [33] have reported several reactions of H_2O_2 reactions with hematin involving O_2 release. But different changes were observed in Alizarin/hemetin/ H_2O_2 system by S. Pirillo *et al.* [5] Our results in the AO7/hemin/ H_2O_2 system indicated that the reactions of oxygen release are also the important reactions in the presence of hemin as catalyst.

In hemin/H₂O₂ system, the O-O bond in H₂O₂ is thought to heterolytic cleavage, giving rise to the active species oxoferryl porphyrin π -cation radical ⁺PFe^{IV}=O. Then the rapid reduction of Fe^{IV} by H₂O₂ happens [33]-[36]:

$$PFe^{III} + H_2O_2 \rightarrow^+ PFe^{IV} = O$$
(3)

$$^{+} PFe^{IV} = O + H_2O_2 \rightarrow [PFe^{III} - O_2^{-}] \rightarrow PFe^{II} + O_2$$

$$\tag{4}$$

Then O_2 may be consumed by the species formed at prior stage or the dye itself. H_2O_2 decomposed easily at basic medium, so the DO is higher in the blank system in which no hemin was added, but the degradation of



Figure 3. Uv-vis spectra changes during the degradation of AO7. [AO] 0.07 mM, $[H_2O_2] 0.09$ mM, $[Hemin] 0.09 \mu$ M, T 15°C.



Figure 4. FTIR spectra changes during the degradation of AO7.



Figure 5. Dissolved oxygen evolution for AO7 degradation process and for Blank at pH 11.0. [AO7] 0.18 mM; $[H_2O_2]$ 0.35 mM; [hemin] 0.27 μ M; T 15°C.

AO7 does not achieve. The results indicate that the oxidative of AO7 by O_2 directly should not be the main reaction.

The main reaction of consuming O_2 is the formation of the superoxide anion [35]:

$$PFe^{II} + O_2 \rightarrow PFe^{III} + O_2^{-}$$
(5)

It has been proposed that the reaction in hematin/ H_2O_2 system is analogous to the classical Fenton reaction and hematin mainly generates •OH radicals in some literatures [5] [19] [33]. In order to check if •OH radicals are involved in AO7 degradation in hemin/ H_2O_2 system, the effects of adding 2-propanol, an •OH radical scavenger [1], on the dye degradation is compared in **Figure 6**.

It is interesting that the degradation rate and and ratio of AO7 is promoted by adding excess 2-propanol under low H_2O_2 concentration. The results indicate that •OH radicals are not the main roles of degradation of AO7 in hemin/ H_2O_2 system in alkaline medium.

 $SO_4^{2^-}$ showed •OH scavenging effect when dye was oxidized by photo-catalyzing and inhibited the degradation of dye, but $SO_3^{2^-}$ has little this effect [37]. But in our studies, 10 mM $SO_4^{2^-}$ had little effect on the degradation of AO7 and 10 mM $SO_3^{2^-}$ greatly inhibited the degradation of AO7. The different effects of $SO_4^{2^-}$ and $SO_3^{2^-}$ in TiO₂/UV and hemin/H₂O₂ systems also illustrate that the •OH radicals are not main roles of degradation of AO7 in hemin/H₂O₂ system.

It has been proposed that protic solvents, including 2-propanol, facilitates O-O bond heterolysis, resulting the high-valence oxoiron(IV) porphyrin π -cation radicals and with the increase of the amounts of alcohol solvents in reaction solutions, the iron porphyrin complex became more electron-deficient[38]. On the other hand, the protic solvents can form H-bond with the –OH in the azo dye, which was benefit for stabilization of azo form [22]. Taking electronic nature of hemin and AO7 structure affected by 2-propanol into considerations, it is assumed that the "PFe^{IV}=O play the important role for degradation of AO7 in hemin/H₂O₂ system.

3.4. Reaction Pathways

To further identify the intermediate products, GC-MS analysis is employed and the intermediate products are identified by comparing the GC spectra and m/z fragments at different reaction time. Taking all the characterizations of degradation analyzed above into consideration, **Figure 7** shows the proposed degradation pathway of AO7 in hemin/H₂O₂ system. The O-O bond in H₂O₂ cleave heterolyticly and the active species "PFe^{IV}=O is produced. The cooperative effects of the π - π and electrostatic interactions between the "PFe^{IV}=O and azo form of AO7 [2] [34] [39] is favor for the electro transfer from the azo chromophoric to the active species of



Figure 6. Effects of 2-propanol on degradation of AO7. [AO] 0.07 mM, [Hemin] 0.09 µM, T 15°C.



Figure 7. Proposed reaction pathway of AO7 degradation in hemin/ H_2O_2 system.

⁺⁺PFe^{IV}=O. Azo chromogenic group is destroyed and benzenesulfonated cation and 2-naphothol radical species form, then they react subsequently with water or superoxide anion to form the products.

4. Conclusions

Oxidative degradation of AO7 with low concentration hydrogen peroxide under hemin catalysis can be achieved in alkaline aqueous solution. Degradation process can be described by a pseudo-first-order kinetic model. The order of dependence on H_2O_2 concentration is significantly larger than that of hemin. Coexisting anions sulphate and chloride have little effect on the degradation, but the reductive sulphite dramatically inhibits the degradation. The protic solvent 2-prophanol obviously promotes the degradation. The azo chromogenic group can be destroyed quickly in hemin/ H_2O_2 system and some smaller intermediates forms during the first degradation of AO7. The active species oxoferryl porphyrin π -cation radical "PFe^{IV}=O generated from heterolytic cleavage of O-O in H_2O_2 catalyzed by hemin plays main roles in degradation of AO7.

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