Analysis of Radium Isotopes in Surface Waters nearby a Phosphate Mining with NORM at Santa Quitéria, Brazil

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ABSTRACT

The radium isotopes $^{226}$Ra and $^{228}$Ra were analyzed in surface water at six points in the neighborhood of a mine of phosphate, associated with uranium, in the region of Santa Quitéria, state of Ceará, Brazil. Water samples were collected during twenty months, filtered and the concentrations of activity determined in the soluble and particulate phases. The results were analyzed using the Principal Component Analysis (PCA) for ordination of environmental data, and also by ANOVA, Tukey and Z tests to compare sets of data considering the radionuclides, the two analyzed phases and the six collecting points. The PCA identified four groups that included all collecting points, using aggregation features such as radionuclide and analyzed phase. The first group is composed by the samples of $^{226}$Ra in the soluble phase; the second group by samples of $^{226}$Ra in the particulate phase; the third one by $^{228}$Ra in the soluble phase, and finally, the fourth group by $^{228}$Ra in the particulate phase. This last group has two discrepant points (01 and 06). Statistical analysis identified differences between the concentrations of activity of radionuclides ($^{228}$Ra higher than $^{226}$Ra) and in analyzed phases (soluble phase higher than the particulate one) but showed no differences between sampled points.

KEYWORDS

Radioecology; Environmental Radioprotection; NORM Mining; Radium Isotopes; PCA

1. Introduction

Phosphate mining and processing can cause significant radiological impacts due to the amount of radionuclides present in the ore [1]. Phosphogypsum e.g., a by-product of phosphate mining, is contaminated by heavy metals and radionuclides, especially $^{226}$Ra [2]. Mining of phosphate with associated Naturally Occurring Radioactive Materials must thus be considered as a NORM activity.

There has been an increased awareness of the radiological impacts of NORM non-nuclear mining pointing that this activity may cause radioactive contamination due to the by-products, wastes and to the installations themselves [3]. In this respect, the environmental impact was analyzed in three practices related to phosphate production: mines, phosphate fertilizers factory and phosphate export platforms. Air particulates, soil, water (lake, river and sea water), biota and plant samples were collected and analyzed. An increase of natural radionuclides in the surroundings of the three enterprises was observed, with fallout being the principal contamination way [4].

Accordingly, the phosphate industry has been recently included within the European regulatory scope. $^{226}$Ra was recognized as the major contaminant. It is found mainly in the processing waters, and the major releasing way is via the liquid effluents [5].

Located in the central-north region of the state of Ceará, Brazil, the “Santa Quitéria” Unit is a phosphate associated with uranium mining installation in predevelopment stage that belongs to the “Indústrias Nucleares do Brasil” (INB). The deposit has recoverable reserves of
about nine million tons of P₂O₅ and 79,500 tons of U₃O₈ [6]. The unit is under the influence of the “Bsh” semiarid climate. The raining season extends from January to May, with sporadic precipitations in June and July. The annual rainfall varies from 550 mm to 960 mm [6]. The rainiest month is March, with rainfall indices varying from 115 mm to 230 mm [6]. Ecologically, the area is characterized as a tropical steppe known as caatinga (savanna) with tropical forests and human occupation areas.

According to Brazilian norms, the unit is classified as a NORM installation [7]. The development of a unit with this classification demands a radiological environmental monitoring program [7-10]. In this context, monitoring means a systematic and planning process of measuring radiation fields, radioactivity and other environmental parameters, including the interpretation of these measurements, in order to characterize, evaluate and control public exposure, especially the critical group, the most exposed to radiation resulting from practice [7-10].

In terms of radioecology, it is necessary to evaluate the behavior of the radionuclides and of their phases in the environment, and how this behavior is able to alter the concentrations of activity of these parameters.

In terms of environmental radiation protection, it is necessary to analyze the possible changes in behavior and in concentrations of activities along the ways of exposition of population and biota in order to evaluate the radiological environmental impact of the project.

A number of radionuclides must be analyzed aiming a comprehensive understanding of their behavior and dispersion in the region. A model for assessment of environmental radiological impact should also be proposed in order to estimate the exposition before the operation and to allow the licensing based in terms of increased dose caused by the practice, as determined by Brazilian laws [2-5].

In Spain, the radiological impact of phosphate mining with NORM decreased significantly with the regulation of the practice and with the separation of the regulation on radiological aspects and chemical aspects that coexist in this kind of practice [11]. The authors also point the extensive regulation of the European Union countries in both aspects of the impact of mining [11].

We here report on the concentrations of activity of the radium isotopes ²²⁶Ra and ²²⁸Ra in surface water measured at six monitoring points in the vicinity of the Santa Quitéria mine. The values were submitted to usual statistical treatment such as the ANalysis Of VAriance (ANOVA), and the Tukey and Z tests [12-15], but also to a data ordination technique used in multivariate statistics known as the Principal Component Analysis (PCA), as an auxiliary tool for the interpretation of measured data [16-18]. The use of univariate and multivariate analyses in the pre-operational environmental monitoring aimed the licensing of the phosphate mine with NORM characteristics, using radium isotopes as a case study.

2. Methodology

2.1. Sampling Area

The Santa Quitéria Unit is located in the municipality of Santa Quitéria, state of Ceará, 212 Km South from the state capital Fortaleza, in northeast of Brazil (Figure 1).

2.2. Sample Collection and Preparation

Environmental water samples (one liter each) were collected monthly during twenty months: in January 2006, and then continuously from June 2006 until December 2007, at six points around the unit, as shown in Table 1. All samples were sent to the Federal University of Ceará where they were filtered through a cellulose acetate filter of porosity 0.45 μm. The fraction that passes through the...
filter was considered as the soluble phase and material retained on the filter was considered as the particulate phase. After filtration, samples were acidified with 1 ml conc. nitric acid per liter and finally sent to the laboratory of environmental analysis of the “Indústrias Nucleares do Brasil” (INB), at the Ore Treatment Unit (UTM) situated at Poços de Caldas, state of Minas Gerais, Brazil, for radionuclides determinations.

2.3. Radionuclides Determination

The determined radionuclides were $^{226}$Ra and $^{228}$Ra. Radium has been chemically separated by co-precipitation from the other radionuclides present in the samples. $^{226}$Ra has been determined by gross alpha counting and $^{228}$Ra has been determined by gross beta counting, as described elsewhere [19].

2.4. Sampling Design and Data Organization

Data were separated by month of collection, collecting point (numbered 01, 02, 03, 04, 05 and 06), radionuclide ($^{226}$Ra or $^{228}$Ra), and sample phase (particulate or soluble phase). Data were organized in a matrix having 24 columns referring to the six collecting points split for each radionuclide and each sample phase, and 20 lines referring to the 20 months of collection. Collected data sum a total of 480 values to be analyzed. Then, the data were grouped in several ways: by radionuclides, regardless of the phase and collecting point, by phase regardless of the radionuclide and the collection point, and finally by collecting point regardless of phase and radionuclide. Thus three factors were analyzed: radionuclide, sample phase and collecting point.

2.5. One Variable Statistic Analysis

The statistical tests used in univariable statistics require adjustments to the Gaussian distribution [12-15]. For this, the adjusting Anderson-Darling test was performed, using the statistical package Minitab® version 16.

An ANOVA was performed to verify the existence of differences in radium concentrations of activity between points, radionuclides and their fractions. This analysis was carried out with the statistical package Excel® version 2010 for Windows® environment. Existing differences the Tukey test was applied to “group” the points with the same concentration of activity.

Four Z tests were made: one to compare the concentrations of activity in the phases (soluble and particulate) for $^{226}$Ra isotope and a second identical for isotope $^{228}$Ra. The third one was carried out between the phases, regardless the radionuclide (i.e. data for the soluble phases for $^{226}$Ra and $^{228}$Ra versus data for the particulate phases for both isotopes). Finally, the fourth Z test was performed with the radionuclides regardless of the phases (i.e. data on $^{226}$Ra in the soluble and particulate phases versus data on $^{228}$Ra in the same phases). These analyzes were done with the statistical package Excel® version 2010 for Windows® environment. Tested hypotheses were:

- $H_0$ there are no differences between means;
- $H_1$ there are differences between means.

2.6. Multi Variable Statistic Analysis

Principal Component Analysis

Principal Component Analysis (PCA) is a technique for modeling co-variaces, which was introduced in 1901 by Pearson [16]. Although multivariate techniques for environmental data assessment are a need, since the world is composed by multi-factorial inter-related parameters, Valentin reported that the first application of PCA in ecology occurred only in 1954 [17].

Nowadays PCA is the most used ordination technique for ecological analysis. Environmental data are ordered in one or two axis. The parameters are established by a correlation similarity of variance-covariance matrix [17]. PCA uses this matrix to produce a set of orthogonal axes ordered from highest to lowest values of a parameter (factor) according to its contribution to the total variance of data. The result is a reduced system of coordinates in which both the position of data in relation to the axes and the relationship between data provide information on the similarities of environmental data [16-18].

3. Results

3.1. Frequency Distribution

Data were analyzed by the Anderson-Darling test for fitness to normal distribution. Concentrations of $^{226}$Ra in the particulate phase gave a test parameter of 15.027 and $^{228}$Ra in the soluble phase a value of 28.064. For $^{228}$Ra, the value in the particulate phase was 8.231 and in the soluble phase 30.422. For all these tests the critical value was less than 0.05. Thus, the four distributions were considered log-normal distributed and therefore a normalizing process was required in the form $y = \ln(x + 1)$, as recommended by Ceteno [12].
3.2. Radium Activity Concentrations

Concentrations of activity of $^{226}$Ra and $^{228}$Ra in water samples, collected during twenty months, were obtained according to described methodology [19]. Table 2 shows the averages of the concentrations of activity and the number of data analyzed for both radionuclides, organized by collecting point and sample phase (before logarithmic transformation). These results can be seen in a graphical way in Figure 2.

3.3. One Variable Analysis

The ANOVA was realized between points, radionuclides and their phases after logarithmic normalizing transformation in the form $y = \ln(x + 1)$ of the values reported in Table 2. In this case, $F_{\text{cal}}$ (value calculated by the test) is equal to 2.28 being higher than $F_{\text{cri}}$ (value if accepted the H$_1$ hypothesis) that is equal to 1.55, with an associated P (statistically significant result) lower than 0.001. This analysis showed differences in the concentrations of activity between the points, radionuclides and phases.

After performing the ANOVA on normalized data, a Tukey test was carried out to detect groups of data with similar average concentrations of activity. Three groups were observed (Table 3).

The first group with the highest mean concentrations of activity appears under label “A” alone (Table 3). This group has only one representative, i.e. $^{228}$Ra in the soluble phase at point 06.

The second group with intermediate concentrations of activity is composed by the 11 factors labeled “A” and “B” simultaneously (Table 3). This group is formed by representatives of $^{226}$Ra and $^{228}$Ra in both soluble and particulate phases, that is $^{228}$Ra in the soluble phase at points 02, 03, 04 and 05, $^{226}$Ra in the particulate phase at points 02, 03, 05 and 06; $^{226}$Ra in the soluble phase at points 01 and 03, and $^{228}$Ra in the particulate phase at point 01.

Finally, the third group with the lowest concentrations of activity is composed by the 12 factors labeled “B” only (Table 3). This group, likewise the second one, is formed by representatives of $^{226}$Ra and $^{228}$Ra in both soluble and particulate phases. This group contains $^{228}$Ra in the soluble phase at points 01 and 04; $^{226}$Ra in the particulate phase at points 02, 04, 05 and 06, and $^{226}$Ra in the particulate phase at points 02, 03, 04, 05 and 06.

Another ANOVA, on normalized data, was performed between collecting points regardless of the phases and radionuclides. This analysis showed the absence of statistical differences between collecting points as $F_{\text{cal}}$ (0.96) was lower than $F_{\text{cri}}$ (2.23), with $P = 0.45$. Thus, all collecting points are considered to have identical means of concentration of activity when the factors radionuclide and phase are not considered.

Comparing among themselves the values of the soluble and particulate phases for each isotope ($^{226}$Ra or $^{228}$Ra), it appears that the mean concentrations of activity of $^{226}$Ra in the phases are considered statistically different ($Z_{\text{cal}} = 2.75 > Z_{\text{cri}} = 1.64$, with $P < 0.01$), with the values in the soluble phase higher than in the particulate one. The same behavior was observed for $^{228}$Ra with...
values of the soluble phase again higher than of the particulate one \( (Z_{cal} = 2.08 > Z_{cri} = 1.64, P < 0.01) \).

When the values of the phases are analyzed regardless of the radionuclides and of the collecting points, the values for the soluble phase were considered statistically higher than those of the particulate phase \( (Z_{cal} = 6.97 > Z_{cri} = 1.64, P < 0.01) \).

Now comparing the radionuclides regardless of the phases and of the collecting points, the concentrations of activity for \( ^{228}\text{Ra} \) were considered higher than those of \( ^{226}\text{Ra} \) \( (Z_{cal} = 2.85 > Z_{cri} = 1.64, \text{with } P < 0.01) \).

### 3.4. Ordination of Data

The PCA analysis results are shown in Figure 3. PCA identified four groups of samples. The results of \( ^{228}\text{Ra} \) in the soluble phase are grouped in the negative parts of axis 1 (Factor 1) and axis 2 (Factor 2).

Results of \( ^{226}\text{Ra} \) in the particulate phase are grouped in the positive part of axis 1 (Factor 1) and in the negative part of axis 2.

Results of \( ^{226}\text{Ra} \) in the soluble phase are grouped along the positive part of axis 1 (Factor 1) and near the origin of axis 2 (Factor 2).

Finally, results of \( ^{228}\text{Ra} \) in the particulate phase are grouped in the positive parts of axis 1 (Factor 1) and axis 2 (Factor 2). This group shows two discrepant points (01 and 06), due to their variance most related to the group composed by \( ^{228}\text{Ra} \) in the soluble phase, mainly point 06.

### 4. Discussion and Conclusions

The statistical analyses allow commentaries. When the radionuclides are analyzed regardless of the other variables, it may be concluded that there are differences between the concentrations of activity of considered radionuclides, with \( ^{228}\text{Ra} \) showing higher concentrations of activity than \( ^{226}\text{Ra} \).

When the phases are analyzed regardless of the other variables, differences are observed between them, with the soluble phase showing higher concentrations of activity than the particulate one.

Concerning the collecting points, another scenario is observed. Thus, when the points are analyzed regardless of the other variables, no differences were observed between them. All the points were considered to have identical means.

Thus, the univariate analysis (ANOVA) allowed establishing that the environmental variables “radionuclide” and “phase” have different behaviors, but the variable “collecting point” showed no differences between the points.

The situation is more complicated when the collecting points are analyzed taking into consideration the variables radionuclide and phase. Three distinct groups appear, with high, intermediate or low means.

The highest average appears for \( ^{228}\text{Ra} \) in the soluble phase at point 06, forming a group of a single element. The second group with intermediate averages is heterogeneous being composed by \( ^{228}\text{Ra} \), in both phases at points 02, 03 and 05, plus \( ^{228}\text{Ra} \) in the soluble phase at point 04 and in the particulate phase at point 06, and also by \( ^{226}\text{Ra} \) in both phases at point 01 and \( ^{228}\text{Ra} \) in the particulate fraction at point 03. The last group, with the lowest concentrations of activity, is composed by \( ^{228}\text{Ra} \) in both phases at point 01 and \( ^{226}\text{Ra} \) in the particulate fraction at point 04. This group also contains \( ^{228}\text{Ra} \) in both phases at points 02, 04, 05 and 06 together with \( ^{226}\text{Ra} \) in the particulate fraction at point 03.
The Tukey test detected the trend of Ra-228 to have higher concentrations than Ra-226, and this was corroborated by the Z test. Thus, Table 3 reports Ra-228 as the only representative of the group of highest activity (A); it is also predominant in the group of intermediate activity AB (8 hits in 11, 73%) but uncommon in the group of lowest activity B (3 in 12, 25%).

In relation to the phase of the radionuclide, values showed no trend. On the contrary, the univariate analysis (Tukey test) identified differences between the collecting points, forming groups of similar activity concentration.

Complementary information was achieved using multivariate analysis. Thus, the PCA technique was used here as a tool to evaluate the possibility of grouping the collecting points depending on the environmental variables radionuclide and phase. Indeed, PCA enabled to ordinate the six collecting points in four groups (Figure 3): one associated the six points with $^{226}$Ra in the soluble phase; a second the six points with $^{226}$Ra in the particulate phase and a third the six points with $^{228}$Ra in soluble phase. Finally, the fourth group associated the points with $^{228}$Ra in the particulate phase. In the latter group, two points (01 and 06) are quite distant producing an irregular group, different from the other three which are nicely homogeneous.

Thus on the whole, PCA was a good method for the ordination of data from monitoring points using concentration of activity of radionuclides and sample phases as parameters.

The combination of univariate and multivariate statistical analyses enabled to assemble a more comprehensive analysis of pre-operational environmental monitoring at the Santa Quitéria phosphate with NORM mine, affording complementary information that only one class of statistical analysis cannot furnish.

Data from this study represent the values of activity concentration in the region before the beginning of the operation of the mine (background). The mine operation inevitably will affect the overall picture presented here and such changes will need to be analyzed in order to assess major concerns relative to radioecology and environmental radioprotection.

REFERENCES


