Trophic Status of Shallow Lakes of La Pampa (Argentina) and Its Relation with the Land Use in the Basin and Nutrient Internal Load

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ABSTRACT

Phosphorus and nitrogen are essential nutrients for living organisms. Their concentration in the water of an aquatic ecosystem is one of the factors responsible for the trophic status of the lake and is related to the soils of the region and to the human activities carried out in their basins. These nutrients are also found in the bottom sediments, where they can either be retained or re-enter the water column. Since the information about the concentrations of nutrients in the water of some lakes of La Pampa (Argentina) is fragmentary, the aim of this study is to describe the trophic status of some shallow lakes of the semiarid center of Argentina and analyze its relation with the human activities in their basins, the concentrations of nutrients and organic matter and particle size distribution of sediments. To this end, we studied ten shallow lakes subjected to different anthropogenic influences (agriculture, agriculture and livestock and impacted by cities). All were hypertrophic and the concentrations of total phosphorus and total nitrogen were among the highest reported globally. Since some lakes had no fish, cladoceran grazing (top-down effect) led them to have reduced concentrations of phytoplankton chlorophyll-a and high water transparency. This relativizes the use of these parameters to determine the trophic status. The sediments of seven of the studied lakes were predominated by fine sands, whereas three were predominated by silts. Nutrient and organic matter content were high, with higher concentrations in lakes with prevalence of fine particles. The reduced adsorption capacity of sediments, the resuspension by wind, the anthropogenic input and the accumulation favored by the arheic character of the basins would explain the high concentrations of nutrients in the water of these Pampean environments.

Keywords: Eutrophication; Trophic State; Total Phosphorus; Shallow Lakes; Internal Load; Land Use

1. Introduction

Phosphorus and nitrogen are essential nutrients for living organisms because they are part of substances of high biological importance such as nucleic acids and proteins. In aquatic ecosystems, these nutrients are taken up by primary producers, either directly from the water in the case of phytoplankton or floating macrophytes or from sediments in the case of benthic algae or rooted macrophytes, where they pass to the rest of the trophic network.

The concentration of nutrients in the water is very variable and phosphorus is one of the factors responsible for the trophic status of an environment. Low concentrations determine oligotrophic ecosystems, with low biological productivity and high water transparency. Conversely, high concentrations determine eutrophic environments, in which the high biological productivity determines low water transparency [1,2]. Thus, as it can be concluded from the above, the trophic status of a water body can also be determined through the analysis of water transparency and the concentration of chlorophyll-a; however, the concentration of phosphorus is one of the strongest predictors of the characteristics of an aquatic ecosystem [3].

The abundance of phosphorus and nitrogen in an aquatic ecosystem is related to the type of soil in the catchment basin, because it tends to be higher in lakes
located in sedimentary basins than in those located in areas with predominance of rocks of volcanic origin [4]. Nutrient concentrations also tend to be higher in environments influenced by human activities that produce discharges, either punctual, as the industrial and domestic ones, or diffuse, as agricultural ones [5-9]. These contributions, which alter the natural cycles, can dramatically change the characteristics of a water body [10,11], and may cause an increase in the biological production, in a process known as eutrophication [1,12].

These nutrients, especially phosphorus, are not only in the water column but also in the bottom sediments, where they can be either retained, forming the internal load [13-15], or resolubilized to re-enter the water column [16-18].

In Argentina, the trophic status of aquatic ecosystems has been studied in lakes at different locations and with contrasting characteristics, such as the deep oligotrophic Patagonian lakes located in the Andes mountain range, south of the 39° or, the hypertrophic shallow lakes of the fertile landscapes of the Pampa Plains, in the province of Buenos Aires, frequently associated with fluvial systems [4,19,20]. Given the high fertility of the soil of this region and the intense agricultural activities carried out in the lake basins, the nutrient levels in these lakes are among the highest cited in previous reports [19,21,22].

In the province of La Pampa, in the semi-arid central region of Argentina, there are a large number of lakes and lagoons, very different from those mentioned above. Their most remarkable characteristics are their scarce depth, for which they can be considered shallow lakes, their temporality and their relatively high and variable salinity. They are located in arheic basins, in a region where rainfalls are overcome by evapotranspiration [23]. It has been determined that the concentrations of nutrients in the water of some of these shallow lakes are very high [24-31], which allows classifying them as hypertrophic [32]. The concentrations of nutrients in the sediments are also very high [27].

Since the information regarding the trophic status of these lakes is partial and fragmentary, the aim of this work is to describe the concentrations of nitrogen and phosphorus in the water and sediments of ten shallow lakes of La Pampa at different locations and under different anthropogenic influence and analyze their relation with the human activities carried out in their basins and with particle size and amount of organic matter in the sediments.

2. Material and Methods

2.1. Study Area

We studied ten water bodies located in different regions of the province of La Pampa (Table 1 and Figure 1).

Six of them, Estancia San José (ESJ), Chadilauquen (Cha), Estancia Pey-Ma (EPM), Ojo de agua Padre Buodo (OaPB), Utracán (Ut) and El Carancho (EC) are semi-permanent lakes subjected to large fluctuations in their surface area and water level. ESJ is located in the northeast of the province, in the region of the Pampa Plains [33] and is influenced mainly by agricultural activities such as cultivation of cereals and oilseeds and livestock rearing. Cha and EPM are located in the Ecotone between the Pampa Plains and the Thorny Forest [33] and are also influenced mainly by agricultural activities. OaPB, Ut, and EC are located in dune areas of the Thorny Forest, where the dominant activity is cattle breeding.

Other three environments Don Tomás (DT), La Arocena (LAr), and Bajo de Giuliani (BG) are located close

<table>
<thead>
<tr>
<th>Lake</th>
<th>Location</th>
<th>Max. length (m)</th>
<th>Max. width (m)</th>
<th>Surface (ha)</th>
<th>Max. depth (m)</th>
</tr>
</thead>
<tbody>
<tr>
<td>1) Chadilauquen (Cha)</td>
<td>64°19’W 35°24’S</td>
<td>6391</td>
<td>2739</td>
<td>893.2</td>
<td>2.2</td>
</tr>
<tr>
<td>2) Estancia Pey-Ma (EPM)</td>
<td>64°15’W 35°26’S</td>
<td>1326</td>
<td>682</td>
<td>62.8</td>
<td>2.1</td>
</tr>
<tr>
<td>3) La Arocena (LAr)</td>
<td>63°42’W 35°40’S</td>
<td>1446</td>
<td>512</td>
<td>48.5</td>
<td>3.2</td>
</tr>
<tr>
<td>4) Estancia San José (ESJ)</td>
<td>63°55’W 36°21’S</td>
<td>2628</td>
<td>1923</td>
<td>288.5</td>
<td>2.47</td>
</tr>
<tr>
<td>5) Don Tomás (DT)</td>
<td>64°19’W 36°37’S</td>
<td>1745</td>
<td>1167</td>
<td>148.3</td>
<td>2.3</td>
</tr>
<tr>
<td>6) Bajo de Giuliani (BG)</td>
<td>64°15’W 36°41’S</td>
<td>7600</td>
<td>2543</td>
<td>1171.3</td>
<td>2.8</td>
</tr>
<tr>
<td>7) Ojo agua P. Buodo (OaPB)</td>
<td>64°17’W 37°14’S</td>
<td>411</td>
<td>165</td>
<td>4.6</td>
<td>2.23</td>
</tr>
<tr>
<td>8) Utracán (Ut)</td>
<td>64°36’W 37°17’S</td>
<td>2333</td>
<td>649</td>
<td>96.6</td>
<td>2.2</td>
</tr>
<tr>
<td>9) El Carancho (EC)</td>
<td>65°03’W 37°27’S</td>
<td>1670</td>
<td>842</td>
<td>85.1</td>
<td>1.6</td>
</tr>
<tr>
<td>10) La Amarga (LAm)</td>
<td>66°07’W 38°13’S</td>
<td>15,428</td>
<td>9688</td>
<td>11,109</td>
<td>12</td>
</tr>
</tbody>
</table>

Table 1. Name, abbreviation, location and main morphometric parameters of the lakes studied.
Trophic Status of Shallow Lakes of La Pampa (Argentina) and Its Relation with the Land Use in the Basin and Nutrient Internal Load

Figure 1. Geographic location of the studied lakes. 1: Chadilaquen; 2: Estancia Pey—Ma; 3: La Arocena; 4: Estancia San José; 5: Don Tomás; 6: Bajo de Giuliani; 7: Ojo de agua Padre Buodo; 8: Utracán; 9: El Carancho y 10: La Amarga. A, B and C: Phytogeographic regions of Pampa Plains, Thorny Forest and Monte respectively (Cabrera, 1976).

to cities and are recipients of storm drains and/or sewage. Due to this significant anthropogenic influence, these originally temporary lakes have become permanent ecosystems. All are shallow lakes with recreational and touristic importance; DT and LAr are adjacent to the cities of Santa Rosa and General Pico, respectively, whereas BG is located 10 km south of Santa Rosa, the capital city of the province of La Pampa.

These nine water bodies are in arheic basins, fed mainly by precipitation and to a lesser extent by phreatic contributions. Water losses occur mainly through evaporation or infiltration.

Finally, we also included La Amarga (LAm) (Figure 1), which differs from the rest by its extent and depth, both of which are considerably higher than those of the rest of the water bodies studied (Table 1), and by its feeding, given mainly by very sporadic contributions from the Chadileuvú River. Given that LAm is located in the phytogeographic province of the Monte [33]; in the most arid region of the province, human activity in the surrounding area is very scarce, restricted to ovine and caprine rearing for subsistence.

2.2. Field and Laboratory Work

Samplings were carried out seasonally during January, April, July and October 2007. Water transparency (Secchi disk) was determined in situ. Water samples were collected at 0.5 m depth with plastic bottles previously conditioned, and refrigerated for physical and chemical analysis and determination of nutrients and chlorophyll-a concentrations.

Sediment samples were collected in November 2007, with an Ekman-type dredge 15 × 15 cm (225 cm²). Samples were placed in polyethylene bags previously conditioned, and kept refrigerated until processing in the laboratory.

The concentration of total Kjeldahl nitrogen (TKN) in water was determined using the method of Kjeldahl and total phosphorus (TP) concentration was determined by digestion of the sample with potassium persulfate in acidic medium and spectrophotometry (Metrolab 1700 spectrophotometer) [34]. The concentration of nutrients in sediments was determined by drying at 55°C - 60°C, after acid digestion, neutralization and spectrophotometric reading with color development by reaction with molybdate in acid medium and reduction with ascorbic acid in the case of TP and with Nessler reagent in the case of total nitrogen.

Chlorophyll-a concentration was estimated by extraction with aqueous acetone with Microclar FFG047WPH filters and spectrophotometry (Metrolab 1700 spectrophotometer) [34,35].

We consider the trophic status classification proposed by OECD [32] (Table 2).

The granulometry of the sediments was determined using a Malvern Mastersizer 2000 particle counter. Car-
bonates were removed with acetic acid 6% and organic matter was removed with 100-volume hydrogen peroxide. The material was dispersed with sodium hexametaphosphate 1% and application of ultrasound for 30 seconds.

To identify the different fractions of the sediment we used the classification shown in Table 3.

A nonparametric Kruskal-Wallis test (H) was applied to determine significant differences between the parameters measured and Spearman correlation coefficients ($r_s$) [36,37] were calculated to examine relationships between environmental factors. To detect possible clustering of lakes, we carried out paired groups and Ward’s method cluster analysis. InfoStat [38] and Past [39] software were used for analyses.

3. Results

Water transparency ranged relatively widely (Figure 2 and Table 4) and was different in the ten lakes studied ($H = 33.53; p = 0.0001$). We recognized two clusters (Figure 3): one including lakes with water transparency greater than 0.7 m and the other including lakes with significantly lower water transparency (Figure 3). Water transparency indicated that the lakes are hypertrophic, with the exception of LAm, which was in the limit of the eutrophy-hypertrophy intervals (Table 2).

The concentration of nutrients in the water of all the lakes was high and since we verified no seasonal pattern, we considered mean values. Regarding TP, EPM showed the highest concentration and was the only environment where the annual mean concentration was higher than 14 mg·L$^{-1}$ (Figure 4 and Table 4). The rest of the lakes showed TP concentrations lower than 10 mg·L$^{-1}$ and OaPB showed the lowest (close to 4 mg·L$^{-1}$) (Figure 4 and Table 4). However, the concentration of TP in water was not different between lakes ($H = 12.53; p = 0.1851$).

Since TP concentrations exceeded widely the lower limit proposed for the category (100 µg·L$^{-1}$), all the water bodies are framed in the category of hypertrophic (Table 2). Regarding TKN, the highest concentration, which exceeded 34 mg·L$^{-1}$, was also recorded in EPM and the lowest, which was close to 8 mg·L$^{-1}$, was re-

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**Table 2. Categories and ranges of trophic status considered.**

<table>
<thead>
<tr>
<th>Trophic Status</th>
<th>Total phosphorus (µg·L$^{-1}$)</th>
<th>Chlorophyll-a (µg·L$^{-1}$)</th>
<th>Depth of Secchi disk (m)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Ultraoligotrophic</td>
<td>&lt;4</td>
<td>&lt;1</td>
<td>&gt;12</td>
</tr>
<tr>
<td>Oligotrophic</td>
<td>4 - 10</td>
<td>1 - 2.5</td>
<td>12 - 6</td>
</tr>
<tr>
<td>Mesotrophic</td>
<td>10 - 35</td>
<td>2.5 - 8</td>
<td>6 - 3</td>
</tr>
<tr>
<td>Eutrophic</td>
<td>35 - 100</td>
<td>8 - 25</td>
<td>3 - 1.5</td>
</tr>
<tr>
<td>Hypertrophic</td>
<td>&gt;100</td>
<td>&gt;25</td>
<td>&lt;1.5</td>
</tr>
</tbody>
</table>

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**Table 3. Dimensions of the particles and size classes.**

<table>
<thead>
<tr>
<th>Fracción</th>
<th>Dimensones</th>
</tr>
</thead>
<tbody>
<tr>
<td>Clay</td>
<td>0.01 - 2 µm</td>
</tr>
<tr>
<td>Silt</td>
<td>2 - 50 µm</td>
</tr>
<tr>
<td>Fine sand</td>
<td>50 - 250 µm</td>
</tr>
<tr>
<td>Medium sand</td>
<td>250 - 500 µm</td>
</tr>
<tr>
<td>Coarse sand</td>
<td>500 - 2000 µm</td>
</tr>
</tbody>
</table>

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**Table 4. Parameters determined in the 10 studied lakes (means and standard deviations).**

<table>
<thead>
<tr>
<th>Lake</th>
<th>TP (mg·L$^{-1}$)</th>
<th>Chlorophyll-a (mg·m$^{-3}$)</th>
<th>Transparency (m)</th>
</tr>
</thead>
<tbody>
<tr>
<td>1) Cha</td>
<td>5.4 (± 4.56)</td>
<td>1.83 (± 1.56)</td>
<td>0.81 (± 0.38)</td>
</tr>
<tr>
<td>2) EPM</td>
<td>14.8 (± 9.76)</td>
<td>14.1 (± 7.81)</td>
<td>0.16 (± 0.11)</td>
</tr>
<tr>
<td>3) LAr</td>
<td>4.5 (± 1.64)</td>
<td>53.63 (± 44.98)</td>
<td>0.42 (± 0.38)</td>
</tr>
<tr>
<td>4) ESJ</td>
<td>5.9 (± 0.63)</td>
<td>5.48 (± 6.83)</td>
<td>0.74 (± 0.28)</td>
</tr>
<tr>
<td>5) DT</td>
<td>4.4 (± 0.88)</td>
<td>145.7 (± 58.4)</td>
<td>0.14 (± 0.04)</td>
</tr>
<tr>
<td>6) BG</td>
<td>7.3 (± 2.07)</td>
<td>204.02 (± 27.5)</td>
<td>0.16 (± 0.05)</td>
</tr>
<tr>
<td>7) OaPB</td>
<td>4.1 (± 0.36)</td>
<td>187.15 (± 62.1)</td>
<td>0.1 (± 0.01)</td>
</tr>
<tr>
<td>8) Ut</td>
<td>9.2 (± 5.64)</td>
<td>1.43 (± 1.39)</td>
<td>1.05 (± 0.39)</td>
</tr>
<tr>
<td>9) EC</td>
<td>6.1 (± 2.0)</td>
<td>1.07 (± 1.67)</td>
<td>1.31 (± 0.42)</td>
</tr>
<tr>
<td>10) LAm</td>
<td>4.4 (± 1.61)</td>
<td>1.65 (± 1.16)</td>
<td>1.54 (± 0.15)</td>
</tr>
</tbody>
</table>

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Figure 2. Mean water transparency of the ten studied lakes.
most 1 mg·m$^{-3}$ to more than 200 mg·m$^{-3}$, so the differences between lakes were significant ($H = 33.33; p = 0.0001$) (Table 4). Cluster analysis allowed observing two clusters: one made up of most of the lakes, with chlorophyll-$a$ concentrations lower than 55 mg·m$^{-3}$ (Figure 5), and the other, due to the high correlation found between chlorophyll-$a$ concentration and water transparency ($r_t = -0.84; p = 0.000$), matching, to a large extent, the one defined based on water transparency, largely composed of DT, OaPB and BG, with concentrations higher than 140 mg·m$^{-3}$ (Figure 5). However, the correlations between chlorophyll-$a$ concentrations and TP ($r_t = -0.13; p = 0.4207$) and TN ($r_t = -0.10; p = 0.5326$) were significant and negative in both cases.

Based on chlorophyll-$a$ concentration, EPM, LAr, DT, BG and OaPB are eutrophic, whereas Cha, ESJ, Ut, EC and LAm were below the limit of 8 mg·m$^{-3}$ proposed for this category (Table 2).

The sediments of seven of the lakes studied (DT, LAr, Cha, OaPB, Ut, EC and LAm) were dominated by middle fraction particles, especially fine sand (Figure 6). In LAr, DT, OaPB, Ut and EC, this type of sediment exceeded 50% of the total. EPM, ESJ, and BG were dominated by thinner fractions, such as silts and clays, exceeding 3% (Figure 6). Despite the short distance between them, Cha and EPM differed in the composition of sediments: the former was dominated by fine sands whereas the latter, as mentioned above, was dominated by silts.

The content of organic matter in the sediments ranged from 0.47% to 9.61% (Figure 7) and was significantly correlated with particle size, since it was always lower in the lakes dominated by sands ($r_t = -0.75; p = 0.0116$) and much higher in the case of the three water bodies dominated by clays ($r_t = 0.75; p = 0.0116$) and silts ($r_t = 0.78; p = 0.0079$).

The TP content in the sediments was high in all the lakes (Figure 8), ranging between 631.8 and 1606.4 mg·kg$^{-1}$. Although it was slightly higher in the lakes whose sediments were dominated by fine particles, we found no significant correlations between both parameters.
The TN content followed a similar pattern, reaching values above 2500 mg·kg$^{-1}$ (Figure 8). We found significant correlations with particle size, since TN content was higher in lakes dominated by clayey ($r_s = 0.79; p = 0.0061$) and silty ($r_s = 0.81; p = 0.0049$) sediments and with higher amount of organic matter. In the lakes dominated by sands, TN was lower than 1900 mg·kg$^{-1}$ and was scarce in the lakes located at the center and southwest of the province (OaPB, Ut, EC and LAm).

### 4. Discussion

The high concentrations of nutrients recorded in the water of the ten shallow Pampean lakes studied in the present work, we found that TP concentrations reached nearly two-fold higher values. Regarding TN, Quirós et al. [21] and Sosnovsky and Quirós [4] reported concentrations close to 28 mg·L$^{-1}$, whereas in the present study we found TKN values higher than 34 mg·L$^{-1}$.

In the province of Buenos Aires, the TP concentration in water is lower in lakes away from urban centers [21]. In contrast, the TP and TKN concentrations in some of the lakes away from urban centers studied in the present work but whose basins are used for agricultural activities tended to be higher than those in lakes influenced by cities, such as DT and LAr. This occurs despite that these lakes receive sewage waste sporadically, and BG permanently receives the waste produced by the nearly 100,000 people living in the city of Santa Rosa. On the other hand, the TP concentration in the water of some Pampean lakes surrounded by crops was found to be slightly lower than that in some lakes located in dune areas surrounded by natural vegetation and whose basins are used for livestock rearing. This could be due to the mechanism of diffuse pollution caused by the dragging of animal feces, especially during a storm, because this is one of the elements with greatest impact on the eutrophication of a water body [5,8], since cattle may excrete between 9 and 16 kg P·ind·year$^{-1}$ [40]. Conversely, TKN concentrations tended to be higher in lakes whose basins are used for agriculture, probably due to the dragging of nitrogen fertilizers.

In some cases, especially in temporary lakes, chlorophyll-$a$ concentrations or water transparency were lower than the limit values proposed by OECD [32] for the category of hypertrophic environments. However, the TP concentrations recorded in the water exceeded that limit widely, which led to negative relations. Thus, in La Pampa, both parameters should be interpreted with caution. This could be due to the presence of zooplanktivorous fish in lakes where there were higher concentrations of chlorophyll-$a$ and therefore lower transparencies (DT, LAr, BG and OaPB). The presence of fish fauna, mainly of the zooplanktivorous silverside (*Odontesthes bonariensis*), is favored by the change from temporary to permanent lakes. Because of the presence of this predator, most of the species found in the zooplankton (especially among cladocerans) are small [19,21,22,28,41-43], a fact that has a great influence on the water quality of the lake. The model of alternative states of shallow lakes indicates that the presence of planktivorous fish in the ecosystem causes a trophic cascade or top-down effect [42,44-46]. The predation of this fish on planktivorous fish in lakes where there were higher concentrations of chlorophyll-$a$ and therefore lower transparencies (DT, LAr, BG and OaPB). The presence of fish fauna, mainly of the zooplanktivorous silverside (*Odontesthes bonariensis*), is favored by the change from temporary to permanent lakes. Because of the presence of this predator, most of the species found in the zooplankton (especially among cladocerans) are small [19,21,22,28,41-43], a fact that has a great influence on the water quality of the lake. The model of alternative states of shallow lakes indicates that the presence of planktivorous fish in the ecosystem causes a trophic cascade or top-down effect [42,44-46]. The predation of this fish on zooplankton of greater size and therefore greater filtration efficiency leads phytoplankton chlorophyll-$a$ concentrations to increase and water transparency to decrease, leading to the state known as turbid [42,44-47].
The bottom sediments of most of the lakes studied were dominated by sands, which is the material that covers most of the surface of the provincial territory [48,49]. The finest sediments were only more abundant in three water bodies, with no regional pattern, given the distance between them and the relatively different landscapes in which they are located.

The sediments of the Pampean lakes have high amounts of nutrients. In some of them, TP concentrations exceeded 1100 mg of P·kg\(^{-1}\), twice more than that verified in hypertrophic lakes of the USA [50] or Spain [51], but were relatively similar to those recorded in shallow lakes in the middle and lower basin of the Yangtze River, in China [52].

The contents of organic matter, TP and TN reported in sediments were higher in the three water bodies that showed highest proportion of silt and clay (BG, EPM and ESJ) and lowest in the lakes dominated by sandy sediments, given the lower adsorption capacity of larger particles [14,53]. Both the nutrients and organic matter of sediments were much more abundant in BG, probably due to the continuous sewage discharges from the city of Santa Rosa received by this lake. The lowest TN concentrations were recorded in the four lakes that are in dune areas or surrounded by native vegetation and where activities in the basin are limited to livestock raising (OaPB, Ut, EC and LAm).

In addition to the influence of the activities carried out in the basins, the high concentrations of nutrients in the waters of the Pampean lakes studied could also be due to a process of internal eutrophication caused by the resolubilization of phosphorus from the sediments and its reentry into the water column [15,54], favored by chemical factors, such as alkalinity [15,54,55] and presence of sulfates [2], physical factors, such as diffusion, bubbling or resuspension of sediments, or biological factors, such as bioturbation [1,17,56]. Among the above-mentioned agents, the resuspension of sediments by the effect of wind is particularly important in shallow lakes, since shallowness favors polymixis [13,16,18,57-59]. It has been verified that, in these lakes, the amount of phosphorus that is resuspended is 5 to 10 times higher than that resuspended in deep lakes [2,60].

5. Conclusions

All the shallow lakes studied in La Pampa were hypertrophic.

The nutrient concentrations in some of the lakes whose basins are used for agricultural activities tended to be higher than those influenced by cities, despite that these lakes receive sewage waste sporadically or permanently. This could be due to the mechanism of diffuse pollution caused by the dragging of animal feces or nitrogen fertilizers.

In some temporary lakes, chlorophyll-\(\alpha\) concentrations and water transparency were lower than the limit for the category of hypertrophic environments, probably due to the absence of planktivorous fish predation, allowing the presence of large zooplankton with high filtration efficiency.

The bottom sediments of most of the lakes studied were dominated by sands and had high amounts of nutrients and organic matter.

A combination of factors such as the reduced absorption capacity of sediments given by the relatively large size of the particles, the frequent resuspension caused by the wind, the continuous income by human activities and the process of accumulation of nutrients favored by the archeic nature of the basin, would explain the high concentrations of phosphorus and nitrogen in the water of the lakes of the province of La Pampa, which are among the highest values reported for lakes in the world.

However, taking into account that some of the lakes studied are temporary and are currently dry because of a period of low rainfall, new studies carried out during different hydroperiods would be useful to compare these results.

REFERENCES


[59] C. Borell Lövstedt and L. Bengtsson, “The Role of Non-
[60] U. Selig and G. Schlungbaum, “Characterisation and

Quantification of Phosphorus Release from Profundal

Battom Sediments in Two Dimictic Lakes during Sum-

http://dx.doi.org/10.4081/jlimnol.2003.151