

Determination of Trace Amounts of Lead by Modified Graphite Furnace Atomic Absorption Spectrometry after Liquid Phase Microextraction with Pyrimidine-2-thiol

Saeid Nazari^{1,2}

¹Department of Chemistry, Faculty of Science, Sabzevar Tarbiat Moallem University, Sabzevar, Iran ²Department of Chemistry, Quchan Branch, Islamic Azad University, Quchan, Iran E-mail: nazari@chemist.com, nazari@sttu.ac.ir Received December 18, 2010; revised February 3, 2011; accepted February 11, 2011

Abstract

The liquid phase microextraction (LPME) was combined with the modified Graphite furnace atomic absorption spectrometry (GF-AAS) for determination of lead in the water and solid samples. In a preconcentration step, lead was extracted from a 2 ml of its aqueous sample in the pH = 5 as lead-Pyrimidine-2-thiol cationic complex into a 4 µl drop of 1,2 dichloroethane and ammonium tetraphenylborate as counter ion immersed in the solution. In the drop, the lead-Pyrimidine-2-thiol ammonium tetraphenylborate ion associated complex was formed. After extraction, the microdrop was retracted and directly transferred into a graphite tube modified by [W.Pd.Mg] (c). Some effective parameters on extraction and complex formation, such as type and volume of organic solvent, pH, concentration of chelating agent and counter ion, extraction time, stirring rate and effect of salt were optimized. Under the optimum conditions, the enrichment factor and recovery were 525% and 94%, respectively. The calibration graph was linear in the range of 0.01 - 12 μ g·L⁻¹ with correlation coefficient of 0.9975 under the optimum conditions of the recommended procedure. The detection limit based on the 3S_b criterion was $0.0072 \ \mu g \cdot L^{-1}$ and relative standard deviation (RSD) for ten replicate measurement of 0.1 μ g·L⁻¹ and 0.4 μ g·L⁻¹ lead was 4.5% and 3.8% respectively. The characteristic concentration was 0.0065 μ g·L⁻¹ equivalent to a characteristic mass of 26 fg. The results for determination of lead in reference materials, spiked tap water and seawater demonstrated the accuracy, recovery and applicability of the presented method.

Keywords: Lead, Liquid Phase Microextraction, Preconcentration, Graphite Furnace Atomic Absorption Spectrometry

1. Introduction

Lead is one of the most ubiquitous elements in the environment and is recognized as a major health risk to humans and animals [1,2]. Lead is a serious cumulative body poison [3] which enters our body system through air, water, and food. Inorganic lead binds itself with the SH group in enzymes or proteins and acts as an enzyme inhibitor [4]. Acute lead poisoning in humans causes severe damage in the kidneys, liver, brain, reproductive system and central nervous system, and even causes death. Mild lead poisoning causes anemia, headache and sore muscles and the victim may feel fatigued and irritable. Chronic exposure to lead causes nephritis, scaring and the shrinking of kidney tissues [5]. It is emitted into the biosphere in considerable amounts, owing to its increased industrial use and its application as a fuel additive [6,7]. In recent years, concern has increased over the concentration of lead in drinking and natural waters [8].

A variety of techniques such as inductively coupled plasma mass spectrometry (ICP-MS) [9], ICP-atomic spectrometry [10,11], electrothermal atomic absorption spectrometry [12] and flame atomic absorption spectrometry (FAAS) [13-16] has been widely used for the determination of trace metal in different samples.

Monitoring trace element concentrations in biological

materials, particularly biological fluids, might be considered a difficult analytical task, mostly due to the complexity of the matrix and the low concentration of these elements, which requires sensitive instrumental techniques and often a preconcentration step. The most widely used techniques for separation and preconcentration of trace lead include liquid-liquid extraction [17], cloud point extraction [18,19], solid phase extraction [20-24] and electrochemical deposition [25], etc. The solvent microextrction technique effectively overcomes these difficulties by reducing the amount of organic solvent and by allowing sample extraction and preconcentration to be done in a single step. The technique is faster and simpler than conventional methods. It is also inexpensive, sensitive and effective for the removal of interfering matrices. Solvent microextraction is a form of solvent extraction with phase ratio values higher than 100 [26-30]. This technique uses simple equipment which is found in most analytical laboratories and also has been used for sample preparation of organic components and has coupled with chromatography methods. We developed this technique in our laboratory and reported for the first time on the coupling of liquid phase microextraction (LPME) with spectrometry to determine inorganic compounds. Using this technique, arsenic in a variety of samples was determined [33].

In this paper, we describe a new and extremely high sensitive method for extraction and determination of lead in aqueous samples by liquid phase microextraction combined with a graphite furnace atomic absorption spectrometry (LPME-GF-AAS). The results indicate that the LPME is an efficient extraction technique for analyzing lead in real samples, with very high pre-concentration factor, greatly increased sensitivity and low detection limit. The method is very simple and quick so that the overall time of extraction and determination for each sample is 8 minutes.

2. Experimental

2.1. Instrumentation

A Shimadzu model AA-6300G atomic absorption spectrometer (Kyoto, Japan) with GFA-EX7i graphite furnace atomizer and D2 lamp for background correction was used. A lead hollow cathode lamp (Hamamatsu photonics, Kyoto, Japan) was used as the radiation source adjusted at the operating current at the wavelength of 283.3 nm with 0.4 nm spectral bandpass. All measurements were performed using peak height and gas stop mode. The measurement conditions are given in **Table 1**. All pH measurements were made by a Metrohm digital pH meter (model: 691, Herisua, Switzerland) with a combined glass electrode.

A 10 μ l Hamilton 7105 syringe (Hamilton, Reno, NV, USA) was used to suspend the drop of the acceptor phase and to inject it into the graphite furnace atomizer. Samples were stirred in 10 ml flat-bottom vial containing Teflon-line septa using an electronic magnetic stirrer (VWR Scientific, West Chester, PA, USA).

Optimum analytical conditions						
Lamp current				4 mA		
Wavelength				228.8 nm		
Spectral bandwidth				0.5 nm		
Spectral bandwidth				peak height		
Purge gas			Ar			
Back ground correct	Back ground correction D ₂			D_2		
GFA heating programme						
Stage	Furnace temperature/°C	Mode	Time/s	Ar flow rate/L·min ^{-1}		
Drying	150	ramp	15	1.5		
Ashing	400	step	15	1.5		
Atomization	1800	step	3	gas stop		
Clean up	2500	step	2	1.5		

Table 1. Applied conditions for lead determination with GFA system.

2.2. Reagents

All reagents used were at least of analytical grade. Water was deionizer to a resistivity of 18 M Ω cm in a Milli-Q system (Millipore, Bedford, MA, USA). A lead stock solution (1000 mg·L⁻¹ Pb) was prepared dissolving high-purity Pb(NO₃)₂ (SPEX, Eddison, NJ, USA).Working solutions were prepared daily in 1% (v/v) HNO₃ by proper dilution of the stock solution. The extraction organic phase was 1,2 dichloroethane (Merck, Darmstadt, Germany). 3.0% (m/v) Pyrimidine-2-thiol (Merck, Darmstadt, Germany) in ethanol was used as a complexing agent and 1.5% (m/v) ammonium tetraphenylborate (Merck, Darmstadt, Germstadt, Germany) in 1,2 dichloroethane was used as counter ion. Modifiers of 0.1% (m/v) Pd in HCl 2% (v/v), 0.1% (m/v) W, 10% (m/v) Mg in water were used.

2.3. Extraction Procedure

Two ml of lead solution was adjusted at pH = 5 and treated with 0.5 ml of 3.0% (m/v) Pyrimidine-2-thiol was transferred to a 10 ml vial. Lead formed a cationic complex with Pyrimidine-2-thiol in aqueous solution. The solution was stirred by magnetic stirrer with a 6 mm stir bar at optimized speed 600 rpm. A 4 µl of 1,2 dichloroethane and ammonium tetraphenylborate as counter ion was taken by the Hamilton syringe whose needle was used to pierce the vial septum. The syringe was clamped in such a way that the tip of the needle was located at a fixed position in the sample solution as shown in Figure **1**. The syringe plunger was depressed to expose the drop and the stirring commenced. In the drop, the Pb-Pyrimidine-2-thiol ammonium tetraphenylborate ion associated complex was formed. The Pb-Pyrimidine-2-thiol cationic complex was extracted from aqueous solution into the 1,2 dichloroethane as extraction organic phase and formed ion associated complex with ammonium tetraphenylborate as counter ion.

After the extraction, the microdrop was retracted and directly injected into the graphite furnace tube modified with [W.Pd.Mg] (c) for subsequent determinations. The different parameters affecting the technique such as solvent, pH, stirring rate, time of extraction, concentration of Pyrimidine-2-thiol and ammonium tetraphenylborate were optimized.

2.4. Tube Modification

Mg modifier was used by injecting 0.2% Mg and sample solution with equal volumes. [(W.Pd) (c) + Pd(i)] modifiers were used for coating containing 40 µg of each of W and Pd from 0.1% of their solutions at temperatures of 2300°C and 2100°C respectively and injecting 10 µl of

0.1% (m/v) Pd on top of 10 μ l sample solution (without extraction). [Pd(c) + Pd(i)] modifier was used as coating of 60 μ g Pd onto the graphite tube at 1800°C and injecting of 10 μ l solution of 0.1% (m/v) Pd on top of 10 μ l of sample solution. [W.Pd.Mg] (c) modifier was used as coating of 40 μ g of each of W, Pd and Mg solution at the appropriate temperatures. Added of pyrimidine-2-thiol to lead solution in direct injection without preconcentration step has not any effect on the signal.

3. Results and Discussion

In order to obtain a high enrichment factor, the effect of different parameters affecting the complex formation and extraction conditions such as type and volume of organic solvent, pH, concentration of chelating agent and counter ion, extraction time and stirring rate were optimized. One variable at a time optimization was used to obtain optimum conditions for liquid phase microextraction (LPME) procedure.

Enrichment factor is defined as the ratio of concentration of lead in the microdrop phase to concentration of lead in the aqueous sample. Concentration of lead in the microdrop phase was calculated from the calibration graph obtained by direct injection of lead into the modified graphite furnace tube without any preconcentration (by compare absorbance of direct injection and after preconcentration). Addition of complexing agent to the lead solution in direct injection without preconcentration step did not any effect on the signal.

3.1. Effect of Type of Modifiers

Several modifiers containing Pd, Ru, Rh, Ir, V, Mo, W, Ni, Mg, Ascorbic acid, separately or in their combinations, were tested. The results of best performing modifiers are shown in **Table 2**. [W.Pd.Mg] (c) modifier showed the best results in contrast to [(W.Pd) (c) + Pd (i)] and [Pd (c) + Pd (i)] for direct determination of lead.



Figure 1. The schematic setup for liquid phase microextraction.

3.2. Effect of Type and Volume of the Extraction Solvent

The choice of organic solvent used in LPME was a major consideration. In order to promote analyte transferring from the donor solution through the organic phase to the accepter microdrop, the solubility of the neutral analytes in the organic solvent should be higher than that in the donor solution and simultaneously the solubility of ionic analytes should be lower than that in the accepter phase. Effect of five different solvents, carbon tetrachloride, dichloromethane, 1,2 dichloroethane, nitrobenzene and benzyl alcohol was evaluated for the of 2 ml lead solution with concentration of 0.1 μ g·L⁻¹. The results are given in **Figure 2**. 1,2 dichloroethane was found to provide higher extraction efficiency. This may be attributed to middle polarity of 1,2 dichloroethane, which leads to

the higher solubility of the polar Pb-Pyrimidine-2-thiol cationic complex and hence higher extraction efficiency.

The influence of drop size was investigated in the range of 1 - 4 μ l. It was found that the absorbance increases with drop volume in the range of 1 - 4 μ l. When drop size exceeded 4 μ l, it became too unstable to be suspended at the needle tip. For this reason, 4 μ l drop volume was used for further studies.

3.3. Effect of Extraction Time

Extraction time is one of the most important factors in the most of extraction procedures. LPME is a type of equilibrium extraction, and the optimal extraction efficiency is obtained when equilibrium is established. Therefore, the extraction time plays a very essential role in the whole process. The dependence of extraction efficiency

Chemical modifier	Detection ^a limit ($\mu g \cdot L^{-1}$)	$\begin{array}{c} \text{Sensitivity}^{\text{b}} \\ (\mu g \cdot L^{-1}) \end{array}$	Linear range $(\mu g \cdot L^{-1})$	RSD % ^c
Mg (i)	0.14	0.038	0.28 - 44	6.2
Ni (i)	0.18	0.042	0.25 - 29	5.8
[(W.Rh) (c) + Rh (i)]	0.11	0.05	0.31 - 38	4.6
$\left[\mathrm{Pd}\left(c\right) +\mathrm{Pd}\left(i\right) \right]$	0.081	0.032	0.30 - 39	5.1
[W.Pd.Mg] (c)	0.074	0.026	0.22 - 3	3.8

a: Based on 3Sb; b: Calculated by dividing 0.0044 to the slope of calibration curve; c: For 10 replicated analysis of 0.1 µg:L⁻¹ lead.

Table 2. Analytical figures of merit for lead determination using different chemical modifiers.



Figure 2. Effect of type of extraction solvent on the absorbance obtained from LPME.

upon extraction time was studied within a range of 0 - 30 minutes in the constant experimental conditions. All measurements were carried out with 0.1 μ g·L⁻¹ lead. **Figure 3** shows the absorbance of lead versus extraction time. The results showed an increase of the lead absorbance up to 8 minutes and leveling off at higher extraction time. Therefore, 8 minutes was used as the optimum extraction time.

3.4. Effect of pH

The effect of pH on the complex formation and extraction of lead from water samples was studied within the range of 3.0 - 9.0. The results, illustrated in **Figure 4**, show that the absorbance is nearly constant in the range of 4.5 - 5.5. In order to obtain high extraction efficiency and minimize diverse ions interferences, pH 5 was chosen.

3.5. Effect of Pyrimidine-2-Thiol Concentration

The influence of the concentration of pyrimidine-2-thiol in the aqueous solution on the lead complex formation was investigated for 0.1 μ g·L⁻¹ solution of lead extracted for 8 minutes. Different concentrations (0.0% - 2.0% m/v)of pyrimidine-2-thiol were used in the aqueous solution and its effects on the extraction process are shown in Figure 5. As can be seen, the efficiency of lead transport increases with increasing pyrimidine-2-thiol concentration until 1.0% (m/v) is reached. However, a further increase in the concentration of pyrimidine-2-thiol (up to 1.0%) caused a pronounced decrease in the formation of lead ion pair. This is most probably due to the competition of pyrimidine-2-thiol itself with lead-pyrimidine-2-thiol complex for transfer through the LPME. Hence, 1.0% (m/v) was employed as the optimum concentration of pyrimidine-2-thiol.



Figure 3. Effect of extraction time on the absorbance of lead obtained from LPME.



Figure 4. Effect of pH on the absorbance of lead obtained from LPME.



Figure 5. Effect of concentration of Pyrimidine-2-thiol on the absorbance of lead obtained from LPME.

3.6. Effect of Ammonium Tetraphenylborate Concentration into the 1,2 Dichloroethane Drop

Different concentrations (0.0% - 1.5% m/v) of ammonium tetraphenylborate as counter ion in the drop of 1,2 dichloroethane under the optimum condition described above were investigated. The results show that by increasing the concentration of ammonium tetraphenylborate the absorbance increases up to 0.8% (m/v) ammonium tetraphenylborate was present in the drop and leveling off at higher concentration as shown in Figure 6.

3.7. Effect of Stirring Rate

Magnetic stirring was used to facilitate the mass transfer process and thus improve the extraction efficiency. The stirring rate was optimized for extraction process. **Figure 7** illustrates the effects of stirring rate on the enrichment factor increased with increasing of the stirring rate up to 600 rpm, because, in high stirring rate, a relatively large vortex is formed in the lower region of the organic solvent, but instability of droplet limited the phenomenon, thus 600 rpm was chosen for further experiment.

3.8. Effect of Salt

The influence of ionic strength was evaluated at 0% - 5% (m/v) NaCl levels while other parameters were kept constant. As observed in **Figure 8**, salt addition has no significantly effect on extraction recovery. Therefore, all the extraction experiments were carried out without adding salt.

3.9. Effect of Foreign Ions

Preconcentration procedures for trace elements in the high salt content samples can be strongly affected by the matrix constituents of the sample. The influence of the common co-existing ions in natural water samples on the lead recovery was investigated. For this purpose, according to the recommended procedure, 2 ml of solution



Figure 6. Effect of concentration of ammonium tetraphenylborate concentration into the 1,2 dichloroethane drop on the absorbance of cadmium obtained from LPME.



Figure 7. Effect of stirring rate on the absorbance of lead obtained from LPME.



Figure 8. Effect of salt concentration on the absorbance of lead obtained from LPME.

that contains 0.1 μ g·L⁻¹ of lead and various amounts from interfering ions, were preconcentrated and determined. A given spices was considered to interfere if it resulted in a ±5% variation of the GFAAS signal. The results are summarized in **Table 3**, proving that the lead recoveries were almost quantitative in the presence of an excessive amount of the possible interfering cations and anions.

Interferent	Concentration ($\mu g \cdot L^{-1}$)	Interferent/Pb (II) ratio	Recovery (%)	
Na ⁺	1,000	10,000	98.6	
Li^+	100	1,000	100.8	
\mathbf{K}^{+}	100	1,000	102.3	
Ca ²⁺	100	1,000	100.1	
Mg^{2+}	100	1,000	99.6	
Ba^{2+}	100	1,000	100.7	
Bi ³⁺	100	1,000	98.2	
Mn ²⁺	100	1,000	97.9	
Co ²⁺	100	1,000	102.1	
Al^{3+}	100	1,000	99.2	
Fe ^{2+ a}	100	1,000	99.7	
Fe ^{3+ a}	100	1,000	99.3	
Ni ²⁺	100	1,000	98.4	
Sn^{4+}	100	1,000	99.1	
Zn^{2+}	100	1,000	100.0	
Cr ³⁺	100	1,000	98.6	
Ag^+	30	300	98.1	
Cd^{2+}	30	300	984	
Cu ²⁺	30	300	99.4	
Si^{4+}	30	300	98.5	
Hg^{2+}	10	100	97.9	
Cl⁻	1,000	10,000	98.2	
Br ⁻	1,000	10,000	98.8	
NO_3^-	100	1,000	101.1	
CH ₃ COO ⁻	100	1,000	99.5	
SCN^-	100	1,000	99.3	
SO_4^{2-}	100	1,000	100.0	
CO_{3}^{2-}	100	1,000	98.6	
PO_{4}^{3-}	100	1,000	99.2	
$S_2O_8^{2-}$	100	1,000	99.7	
SeO_3^{2-}	100	1,000	98.4	

Table 3. Effect of interferents on the recovery of 0.1 µg·L⁻¹Pb(II) in water sample using LPME.

^aMasked with F⁻.

3.10. Analytical Figure of Merits

Under the optimum conditions described above, the analytical performance characteristics of the proposed method are listed in **Table 4**.

The calibration graph, obtained by analysis of 8 standards with different known concentrations of Pb, was linear with a correlation coefficient of 0.9996 at levels near the detection limits and up to at least 4.5 μ g·L⁻¹ [A = 0.0021 (±0.002) + 0.25068 (±0.0021)C], where A is the absorbance and C is Pb2+ concentration in μ g·L⁻¹.

The calibration graph was linear in the range of 0.01 -

12 $\mu g \cdot L^{-1}$ with correlation coefficient of 0.9975 under the optimum conditions of the recommended procedure. The equation of line is A = 0.0045 (±0.0062) + 0.48C (±0.0085), where A is the absorbance and C is concentration of lead in $\mu g \cdot L^{-1}$ in the initial solution. The detection limit was calculated as three times the standard deviation of the peak absorbance for injection of 4 μ l of ten extractions of the blank, using the liquid phase microextraction procedure. The detection limit was calculated to be 0.0072 $\mu g \cdot L^{-1}$ with absolute value of 28.8 fg for 4 μ l injection into the graphite furnace. The characteristic concentration was 0.0065 $\mu g \cdot L^{-1}$ equivalent to a characteristic mass of 26 fg. The relative standard deviation for (RSD) for ten replicate measurement of 0.1 μ g·L⁻¹ and 0.4 μ g·L⁻¹ lead was 4.5% and 3.8% respectively. The enrichment factor (EF) was obtained from the slope ratio of calibration graph after and before extraction, which was about 525. The extraction recovery (R%) was 94% which was calculated by Equation (1).

$$R\% = (Vdrop/Vsolution) \times EF \times 100$$
 (1)

3.11. Application

In order to establish the validity of the proposed procedure, the method has been applied to the determination of lead in standard reference materials NIST 1643e, NIST 1640 (National Institute of Standard and Technology NIST, USA), JB-1, JB-1a and JB-2 as powder obtained from geological survey of Japan (GSJ). The powder was dissolved in 15 ml of a mixture of 500 ml HF, 165 ml H₂SO₄ and 40 ml HNO₃ at 150°C in a teflon beaker overnight. To show the applicability of the method, seawater from Caspian Sea and local tap water was analyzed for its lead content. The results obtained are presented in Table 5. As shown by the results in Table 5, good agreement between certified and found values was obtained at a 95% confidence level, indicating that calibration carried out using aqueous standard solutions submitted to the LPME procedure results in good accuracy. To assure the homogeneity and statistical validity of the method, a paired *t*-test was applied to the group of results; the determined "t" value was 0.187, which is below the reference t-value for a 95% confidence interval (t = 2.78).

Table 4. Analytical characteristics of LPME-GF AAS for determination of lead.

Parameter	Analytical feature	
Linear range ($\mu g \cdot L^{-1}$)	0.01 - 12	
Correlation coefficient (r ²)	0.9975	
Limit of detection $(\mu g \cdot L^{-1})(3\sigma, n = 10)$	0.0072	
Repeatability (RSD %) (n = 10, 0.1 μ g·L ⁻¹)	4.5	
Repeatability (RSD %) (n = 10, 0.4 μ g·L ⁻¹)	3.8	
Enrichment factor (EF) ^a	525	
Sample Volume (ml)	2	
Sample introduction Volume (µl)	4	
Sample preparation time (min)	8	
Recovery (%)	94	

^aEnrichment factor is the slope ratio of calibration graph after and before extraction.

Table 5. Results obtained for the determination of Pb in SRM using LPME-GFAAS with a modified graphite tube (n = 3; *t*-Student applied for 95% confidence level; t = 0.1874 for the group of results).

Sample	Lead added $(\mu g \cdot L^{-1})$	Lead found $(\mu g \cdot L^{-1})$	Recovery %
Sea water*	0	3.7 ± 0.1	_
	0.2	0.196 ± 0.01	98.3
	0.4	0.39 ± 0.02	97.4
	0.6	0.59 ± 0.04	99.2
Tap water	0	0.00	
	0.2	0.202 ± 0.01	101
	0.4	0.41 ± 0.03	102.5
Reference Material	Certified value	Measured value	Recovery %
JB-1 ^a	$10.0 \ (\mu g \cdot g^{-1})$	$9.93 \pm 0.41 \ (\ \mu g \cdot g^{-1})$	99.3
JB-1a ^a	6.76 (µg·g ⁻¹)	$6.66\pm 0.32\;(\;\mu g{\cdot}g^{-1})$	99.0
JB-2 ^a	$5.36 (\mu g \cdot g^{-1})$	$5.19 \pm 0.26 \text{ (} \mu g \cdot g^{-1}\text{)}$	96.9
NIST 1643e ^b	$19.63 \pm 0.21 \; (\mu g \cdot L^{-1})$	$20.07 \pm 0.84 \; (\mu g \cdot L^{-1})$	102.2
NIST 1640 ^b	$27.89 \pm 0.14 \ (\mu g \cdot L^{-1})$	$28.84 \pm 1.4 \ (\mu g \cdot L^{-1})$	103.4

*Collected from Caspian Sea; *Obtained from geological survey of Japan, GSJ; *From National Institute of Standard and Technology NIST (USA).

Method	$\begin{array}{c} LOD \\ (\mu g \cdot L^{-l}) \end{array}$	RSD (%)	Enrichment factor	Sample volume (ml)	Sample preparation time (min)	Reference
Liquid–liquid extraction and micro volume back-extraction–FAAS	0.39	6.3	543	500.0	> 6	31
Co-precipitation-FAAS	16	3.0	125	50.0	> 20	32
Off-line-SPE-FAAS	6.1	4.7	30	300	4	33
On-line-SPE-FAAS	0.8	2.6	330	39	4	34
CPE-FAAS	1.1	3.5	50	50.0	30	35
LPME-ETAAS	0.0072	3.8 ^a , 4.5 ^b	525	2	8	Represented method

Table 6. Characteristics performance data obtained by using LPME and other preconcentration techniques for determination of lead.

^aLead concentration was 0.1 μ g·L⁻¹ for which RSD was obtained; ^bLead concentration was 0.4 μ g·L⁻¹ for which RSD was obtained.

3.12. Comparison with Other Methods

Table 6 indicates the limit of detection (LOD), the relative standard deviation, the sample preparation time and the sample volume in the LLE-microvolume back-extraction [31], co-precipitation [32], off-line SPE (Solid Phase Extraction) [33], on-line SPE [34] CPE and (Cloud Point Extraction) [35] for the extraction and determination of lead in water samples. The comparison of the results exhibits that LOD and the enrichment factor (or enhancement factor) in the present method were better than those of the other methods.

4. Conclusions

The results show a very promising technique for the determination of lead in variety of samples at $\mu g \cdot L^{-1}$ levels without the needs for any sophisticated device. Apart from extremely high sensitivity and relatively free from interferences, the procedure is very simple, fast and benefits a very low detection limit. The method is also inexpensive and reproducible and applied for sea and tap water samples. The method can also be applied for analysis of real samples such as biological and botanical samples.

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